

Influence of mediated processes on the removal of Rhodamine with CDEO

Danyelle Medeiros de Araújo¹, Cristina Sáez², Carlos A. Martínez-Huitle^{1,*},
Pablo Cañizares², M. A. Rodrigo²

¹ Institute of Chemistry, Federal University of Rio Grande do Norte, Lagoa Nova CEP
59078-970, Natal, RN, Brazil

²Department of Chemical Engineering. Enrique Costa Building. Campus Universitario
s/n. 13071 Ciudad Real. Spain

Abstract

The influence of the mediated oxidation on the removal of Rhodamine B (xanthene dye) solutions with conductive-diamond electrochemical oxidation (CDEO) is studied. To do this, four different supporting electrolytes have been used: Na₂SO₄, HClO₄, H₃PO₄, NaCl. Total removal of organic pollutants is attained with CDEO regardless of the supporting electrolyte media used, although media clearly influences on efficiency and rate of the processes. Sulfate and phosphate media show a similar behavior, whereas electrolysis in perchlorate media behaves surprisingly better than chloride media. Current density is playing an important role. In all cases, CDEO follows a first order kinetic (linear trend in semi logarithmic plot) and kinetic constants are generally much greater than expected according to a single mass transfer electrolytic model. This is not the expected result for a direct electrochemical oxidation process and it indicates the importance of mediated electrochemical processes in the removal of Rhodamine B. The harsh oxidation conditions of CDEO leads to the formation of less reaction

26 intermediates than other technologies. The presence of short chain aliphatic acids is
27 discharged, and the intermediates (aromatic acids) formed during the initial stages of the
28 process are rapidly mineralized to carbon dioxide. In chloride media, chlorinated
29 intermediates are also formed by the action of hypochlorite.

30

31 *Key Words:* Electrochemical oxidation, diamond, dyes, mediated oxidation,
32 supporting electrolyte.

33

34 * Corresponding author:

35

36

37

38

39

40

41

42

43

44

45

46

47

48 **1. Introduction**

49 The treatment of textile effluents has given much attention in the last years because the
50 discharge of highly pigmented synthetic dyes to the ecosystem causes significant
51 damage to aquatic and human lives (Carneiro et al., 2010). Conventional technologies
52 are inefficient to remove these organics due to the high water solubility of dyes (in the
53 case of using coagulation and precipitation (Muthukumar et al., 2005)), and to the
54 typical pH and salt concentration of dye effluents (in the case of using biological
55 methods (Pandey et al., 2007)). In this point, Advanced Oxidation Processes appears as a
56 good alternative for the treatment of synthetic dyes effluents. Among them, Fenton
57 oxidation (Cruz-Gonzalez et al., 2012; Panizza and Cerisola, 2009b), photocatalysis (He
58 et al., 2009; Miller et al., 2013), sonochemical degradation (Guzman-Duque et al., 2011;
59 Merouani et al., 2010) and electrochemical oxidation (Martinez-Huitle et al., 2012;
60 Rocha et al., 2012; Saez et al., 2007; Tavares et al., 2012) have been widely studied in
61 literature for the treatment of a great variety of dyes. In general, these technologies are
62 able to attain very good results in terms of decolorization due to the rapid cleavage of
63 chromophore group of the dye molecule. However, mineralization efficiency shows a
64 marked influence of the oxidative capacity of each degradation technology. Generally,
65 aromatic and aliphatic acids are the main intermediates detected during dye degradation
66 process, although their maximum concentration depends on the conditions used in each
67 case. In this point, recent works have shown that electrochemical oxidation, and in
68 particular Conductive Diamond Electrochemical Oxidation (CDEO), can be
69 successfully applied with high organic removal rates and without important operational
70 limitations (Aquino et al., 2013; Aquino et al., 2012; Faouzi et al., 2006; Ramirez et al.,
71 2013; Saez et al., 2007). The harsh oxidation conditions attained with CDEO are
72 explained in terms of the oxidation mechanisms involved in the process. In this way, it

73 is well documented that during CDEO a noteworthy production of hydroxyl radicals in
74 the nearness of diamond surfaces takes place. These radicals are fully available for
75 oxidizing reactions but their lifetime is very short and not enough to let them diffuse to
76 the bulk solution. Thus, the action of these radicals is limited to the short region close to
77 the electrode surface where they are produced, and the kinetic of these processes is
78 usually controlled by mass transport. Additionally, it is well documented the availability
79 of CDEO to produce inorganic oxidants, which are difficult or even impossible to be
80 produced with other different anodic materials, from the oxidation of supporting
81 electrolyte (Cañizares et al., 2005a; Cañizares et al., 2009; Sáez et al., 2008). In fact,
82 some works have also been focused on the electrochemical synthesis with diamond
83 anodes of powerful oxidants such as persulfates (Serrano et al., 2002), perphosphates
84 (Cañizares et al., 2005a), perchlorates (Bergmann et al., 2014; Sánchez-Carretero et al.,
85 2011) and hypochlorite (Cañizares et al., 2009; Vacca et al., 2011). Thus, besides direct
86 electrooxidation on the surface and oxidation by means of hydroxyl radicals in a region
87 close to the electrode surface, the oxidation mediated by other oxidants electrogenerated
88 on the diamond surface from the electrolyte salts should be taken into account, as it
89 seems to increase the global oxidation efficiency (Aquino et al., 2012; Cañizares et al.,
90 2005b; Panizza and Cerisola, 2009a; Rodrigo et al., 2010).

91 The dye under consideration is RhB. It is widely used in textiles, leathers and food
92 stuffs with high water solubility (Hou et al., 2011; Miller et al., 2013). Thus, this work
93 focuses on the CDEO of synthetic Rhodamine B (selected as model of xanthene dyes
94 with very good stability) solutions in different supporting media in order to increase the
95 understanding of the role of mediated oxidation on the degradation process.

96

97 **2. Materials and methods**

98 **2.1. Chemicals**

99 Rhodamine was supplied by Sigma-Aldrich Laborchemikalien GmbH (Steinheim,
100 Germany). Anhydrous sodium sulphate, sodium chloride, perchloric acid and
101 phosphoric acid, used as supporting electrolytes, were analytical grade purchased from
102 Fluka. All solutions were prepared with high-purity water obtained from a Millipore
103 Milli-Q system, with resistivity > 18 MΩ cm at 25 °C. Sulphuric acid and sodium
104 hydroxide used to adjust the solution pH were analytical grade and supplied by Panreac
105 Química S.A. (Barcelona, Spain).

106

107 **2.2. Analytical procedures**

108 The Total Organic Carbon concentration was monitored using a Multi N/C 3100
109 Analytik Jena analyzer. Measurements of pH and conductivity were carried out with an
110 InoLab WTW pH-meter and a GLP 31 Crison conductimeter, respectively. The
111 concentrations of the compounds were quantified by HPLC (Agilent 1100 series). The
112 detection wavelength used to detect dye was 248 nm. The column temperature was 25
113 °C. Volume injection was set to 50 µL. The analytical column used was Phenomenex
114 Gemini 5 µm C18. The mobile phases were 0.9 ml of 98% formic acid in milli-Q water
115 for analysing intermediates. Samples extracted from electrolyzed solutions were filtered
116 with 0.20 µm Nylon filters before analysis. Moreover, the acids intermediates formed
117 during the experiments were detected with a detection wavelength of 190 nm.

118 Samples of anolyte were extracted into non-aqueous medium (2 mL of acetonitrile
119 HPLC grade with 20 µL of electrolysis sample) and were subjected to GC-MS analysis
120 using GC-FOCUS and MS-ISQ Thermo Scientific to identify the intermediates
121 following the conditions: GC: Varian column VF5 ms with a composition of 5% de
122 fenil-arylene and 95% de dimetilpolisiloxane. Temperature program: 40°C – 5 min;

123 12°C/min – 100°C; 10°C/min – 200 °C and 10°C/min - 270 °C – 5 min. Injector: 220°C.
124 Mode: Splitless. Gas flow: 0.8 mL/min. MS: Transfer line: 270°C; ions source
125 temperature: 220°C, Mass range: 40-500 m/z. Injection: 1µL.

126

127 **2.3. Electrochemical cells**

128 Electrolyses were carried out in a single compartment electrochemical flow cell
129 working under a batch-operation mode (Cañizares et al., 2005b). Conductive –Diamond
130 Electrodes (p-Si–boron-doped diamond) were used as anode and a stainless steel (AISI
131 304) as cathode. Both electrodes were circular (100 mm diameter) with a geometric area
132 of 78 cm² and an electrode gap of 9 mm. Boron-doped diamond films were provided by
133 Adamant Technologies (Neuchatel, Switzerland) and synthesized by the hot filament
134 chemical vapour deposition technique (HF CVD) on single-crystal p-type Si <100>
135 wafers (0.1 Ωcm, Siltronix).

136

137 **2.4. Experimental procedures**

138 Bench-scale electrolyses of 1000 cm³ of wastewater were carried out under
139 galvanostatic conditions. The current density employed ranged from 15-120 mA cm⁻².
140 The cell voltage did not vary during electrolysis, indicating that conductive-diamond
141 layers did not undergo appreciable deterioration or passivation phenomena. Prior to use
142 in galvanostatic electrolysis assays, the electrode was polarized during 10 min in a
143 0.035 M Na₂SO₄ solution at 15 mA cm⁻² to remove any kind of impurity from its
144 surface.

145 The wastewater consisting of 100 mg dm⁻³ of Rhodamine B and different electrolytes
146 (Na₂SO₄, HClO₄, H₃PO₄, NaCl) was stored in a glass tank and circulated through the
147 electrolytic cell by means of a centrifugal pump (flow rate 21.4 dm³ h⁻¹). A heat

148 exchanger coupled with a controlled thermostatic bath (Digiterm 100, JP Selecta,
149 Barcelona, Spain) was used to maintain the temperature at the desired set point (25 °C).

150

151 **3. Results and discussion**

152 Figure 1 shows the progress of the mineralization during electrolyses at two large
153 current densities (60 and 120 mA cm⁻²) of synthetic wastewater consisting of aqueous
154 solutions containing 100 mg dm⁻³ of Rhodamine B (RhB) and different electrolytes
155 (Na₂SO₄, HClO₄, H₃PO₄, NaCl). As it can be observed, total removal of organic
156 pollutants can be attained with CDEO regardless of the supporting electrolyte media
157 used, although media clearly influences on efficiency and rate of the processes, and
158 current density is playing an important role. This is interesting because it confirms that
159 CDEO is robust enough to deplete pollution but it also informs about the occurrence of
160 different mechanisms of oxidation that should be related with the supporting media.
161 Thus, although chloride media could be expected to show the best results because of the
162 well-known production of chlorine (and then of hypochlorite and hypochloric acid), it
163 attains the worst results and, in fact, it is the only case in which the complete
164 mineralization is not attained for a current charge applied of 100 Ah dm⁻³. This behavior
165 can be attributed to an interaction between hydroxyl radicals and Cl⁻ (chloride acts as
166 scavenger of •OH) to form different active chlorine species on BDD surface (Bergmann
167 et al., 2009; Cañizares et al., 2009; Sánchez-Carretero et al., 2011) with this non-active
168 material and to the formation of refractory species by chlorination. At the same time, the
169 chloride concentration in the solution can also promote the importance of Cl₂
170 production at BDD surface, decreasing the Cl⁻, and consequently, the production of
171 active chlorine species.

172 A very interesting observation is the effect of current density in the perchlorate test that
173 shifts towards higher efficiencies at higher current densities. Initially, it is strange
174 because in that media no oxidants production is expected and perchlorate is not a good
175 oxidant at room temperature so, it is not expected to participate in the oxidation.
176 However, higher concentration of hydroxyl radicals are produced under these
177 experimental conditions as well as it can be related to the nature of the hydroxyl radicals
178 produced at different non-active anodes, as already proposed by Bejan and co-workers
179 (Bejan et al., 2012). On the other hand, sulfate and phosphate media show a similar
180 behavior in both current densities tested.

181 Figure 2 shows a semi-logarithmic plot of the COD decay with the current charge for
182 the same current densities shown in Figure 1. As it can be observed, except for an
183 abrupt decrease during the very first stages of the electrolysis of the synthetic
184 wastewater with chloride media, all tests follow a clear linear trend suggesting that they
185 can be modeled with a first order kinetic, as they were carried out under galvanostatic
186 conditions. This divergence can be explained in terms of the masking effects of
187 chlorides on the measurement of COD. Not great differences are obtained if results of
188 the electrolyses of sulfate and phosphate media are compared. In fact, results with these
189 two supporting electrolytes do not seem to depend on the current density and on the
190 applied charge. However, as it was forwarded by the changes in the mineralization,
191 perchlorate removal rate increases very importantly when current density increases. The
192 same can be observed in the electrolysis in chloride media.

193

194 To better observe this influence of the supporting media and current density, and taking
195 into account the linear trend observed in COD decay vs. Q , experimental results of each
196 test were fitted to a first order kinetic decay. Figure 3 shows the influence of the

197 supporting media and applied current density on the value of these first order kinetic
198 constants for COD removal. These constants were calculated from the slope of semi-
199 logarithmic vs. time plots.

200

201 For the case of NaCl, the points corresponding to the first oxidation stage (which fit
202 well to this trend) were used to fit the constant, whereas for the rest of supporting
203 electrolytes the complete set of data was used. The increase of kinetic constants with the
204 applied current density is more important from current densities above 100 mA cm^{-2} .
205 This is the expected behavior of a process with a significant contribution of mediated
206 electrooxidation: the higher the current density, the higher the production of mediated
207 electro-reagents and, hence, the higher the kinetic constant. The lower values obtained
208 in the electrolysis carried out with sodium chloride as supporting electrolyte can be
209 related to the importance of Cl_2 production at BDD surface, decreasing the production
210 of active chlorine species, and consequently, disfavoring the oxidation of organic matter.
211 In addition, it is important to notice that calculated kinetic constants are much greater
212 than the value expected for a purely mass transfer controlled process (around 0.01 min^{-1}
213 ¹), mainly in the case of working at 120 mA cm^{-2} . This value was calculated from a
214 ferro-ferricyanide standard test carried out with the same experimental setup and under
215 the same flow conditions (Aquino et al., 2012). This certifies that mediated oxidation
216 carried out by electrogenerated oxidants in bulk region plays an important role in the
217 degradation of RhB and mineralization process.

218

219 Efficiency of electrolytic processes can be discussed in terms of different parameters
220 such as ICE, ACE, etc. However, in case of a good fitting of results in galvanostatic
221 experiments to a first order kinetics (as it is the case of the results obtained in this work),

222 it can be demonstrated that efficiency depends linearly on the COD concentration and
223 the constant can be easily calculated from the kinetic constant as reported in literature
224 (Aquino et al., 2014). This constant is of a great importance because it gives directly the
225 value of the efficiency (% of the applied current used in oxidizing organics) if
226 multiplied by the COD (mg dm^{-3}) and in the authors' opinion is the most interesting
227 efficiency parameter to characterize CDEO processes. Figure 4 shows the changes in
228 this parameter for the four supporting electrolytes and current densities assessed. As it
229 can be observed, efficiency decreases as the current density increases, although it
230 increases slightly at high current density (above 100 mAcm^{-2}).

231

232 Other important point to be studied is the production of intermediates, because some of
233 them can be even more hazardous than the initial pollutant. Their relative concentration
234 and trend can help to increase the understanding of the process. Figure 5 shows the
235 chromatographic area of the main intermediates obtained during the electrolyses at 60 mA
236 cm^{-2} of RhB solutions containing different supporting electrolytes. The chromatographic
237 area of the peak corresponding to RhB has been also included for the sake of
238 comparison. As it can be observed, the harsh oxidation conditions of electrochemical
239 oxidation do not favor the production of many intermediates (only three intermediates
240 seems to be formed in significant concentration), except in the case of NaCl media in
241 which a very rapid depletion of RhB molecule is observed together with the production
242 of seven compounds that clearly behaves as intermediates. Their formation can be
243 related to the attack of ClO^- electrochemically generated by oxidation of Cl^- to RhB
244 molecule. This observation can explain the trend observed in COD decay during the
245 electrolysis of RhB in chloride media. However, in all cases, intermediate
246 concentrations should be very small as compared to that of RhB. As an indication, the

247 chromatographic area of the RhB peak (with a concentration of 100 mg dm^{-3}) was
248 around 170.000 units, while the maximum response measured for intermediates during
249 the experiments is below 25.000 units for i_2 and below 5.000 units for the rest of
250 intermediates.

251 Restricting now our analysis to the intermediates produced and identified, a lower
252 amount of intermediates (Phthalic acid (\blacktriangle); 2,5-Hydroxybenzoic acid (\bullet) and Benzoic
253 acid (\blacksquare), with low-units of chromatographic area) is produced when HClO_4 is used as
254 supporting electrolyte, being completely oxidized after 300 min of electrolysis (Fig. 5a).
255 For the case of H_3PO_4 and Na_2SO_4 , the same intermediates were identified (Figs. 5a, 5c
256 and 5d); however, the concentration of them is very different than those obtained to
257 HClO_4 . An important observation is the significant concentration produced of 2,5-
258 hydroxybenzoic acid when HClO_4 , H_3PO_4 and Na_2SO_4 are used as supporting
259 electrolytes (Fig. 5a, 5c and 5d) while the concentration of the other intermediates is
260 minor (phthalic acid (\blacktriangle) and benzoic acid (\blacksquare), respectively). However, the production
261 of 2,5-hydroxybenzoic acid indicates that RhB is fragmented to form benzoic acid,
262 which is successively oxidized to 2,5-hydroxybenzoic acid, justifying its higher
263 concentration. Then, RhB is preferentially oxidized by strong oxidants (hydroxyl
264 radicals and persulfates/ peroxodiphosphates) produced at BDD surface. These
265 intermediates are completely eliminated after 300 min, in all cases.

266 On contrary, for NaCl , seven sub-products are produced (phthalic acid (\blacktriangle); 2,5-
267 hydroxybenzoic acid (\bullet); benzoic acid (\blacksquare); 3-dinitrobenzoic acid (\blacklozenge); α -hydroxyglutaric
268 acid (\triangle); intermediate 6 (\circ) and intermediate 7 (\square)) when RhB was electrochemically
269 oxidized. No complete elimination of all intermediates formed was accomplished after
270 500 min of electrolysis, but lower concentration of them remains in solution as showed

271 by minor chromatographic areas recorded. This fact may be explained by the effective
272 attack of hydroxyl radicals and active chlorine species produced at NaCl media to RhB.

273 Figure 6 shows the maximum relative chromatographic area of the intermediates
274 generated at different applied current density and supporting media. Comparing results,
275 neither current density nor supporting media influence significantly in the maximum
276 concentration of each intermediate, except in the case of chlorine media in which
277 chlorinated intermediates in lower relative concentration seem to be produced
278 (intermediate 6 (○) and intermediate 7 (□)). At the light of these results, CDEO leads to
279 the formation of less reaction intermediates than other technologies such as
280 photocatalysis degradation (He et al., 2009; Yu et al., 2009; Sun et al., 2009) and Fenton
281 oxidation (Hou et al., 2011) in which benzoic and phthalic acids, aliphatic acids (mainly
282 adipic and glutaric acids) and alcohols of short chain are formed in relevant
283 concentrations. In this case, short chain intermediates are not detected (at least with the
284 analytical method used). The strong oxidation conditions of this technology favor the
285 mineralization of principal intermediates into carbon dioxide and refractory species are
286 not formed in any case.

287

288 **4. Conclusions**

289

290 From this work the following conclusions can be drawn:

291

- 292 – RhB can be successfully removed using conductive-diamond electrochemical
293 oxidation regardless of the supporting electrolyte media, although media clearly
294 influences on efficiency and rate of the processes, and current density is playing
295 an important role.

296

297 – Kinetic constants are much greater than the value expected for a purely mass
298 transfer controlled process, indicating that there is a significant contribution of
299 mediated oxidation processes. Current density promotes the formation of
300 mediators and rate of the electrolysis improves with increasing this parameter.
301 Opposite efficiency decreases with an increase in current density.

302

303 – The harsh oxidation conditions of CDEO favor the rapid mineralization of RhB
304 into carbon dioxide. Aromatics intermediates (mainly phthalic acid and benzoic
305 acids) formed by the cleavage of RhB molecule are the only intermediates
306 detected in the first stages of the degradation process. Neither short chain
307 intermediates nor refractory species are formed in any case. In case of chloride
308 media, chlorinated intermediates can be formed by the attack of hypochlorite to
309 organic compounds.

310

311 **Acknowledgements**

312 D.M.A. acknowledges CAPES for PhD fellowship and CNPq for the scholarship given
313 for “doutorado sanduiche” under “Ciências sem Fronteiras” program to develop the
314 experimental research at the UCLM-Spain. This work has been supported by the
315 Spanish Government through the project CTM2013-45612-R and EU through project
316 FEDER 2007-2013 PP201010 (Planta Piloto de Estación de Regeneración de Aguas
317 Depuradas).

318

319 **References**

- 320 Aquino, J. M., R. C. Rocha-Filho, M. A. Rodrigo, C. Sáez, and P. Cañizares, 2014, High efficiencies
321 in the electrochemical oxidation of an anthraquinonic dye with conductive-diamond
322 anodes, *Environmental Science Pollution Research*, DOI 10.1007/s11356-014-2784-0.
- 323 Aquino, J. M., R. C. Rocha-Filho, M. A. Rodrigo, C. Sáez, and P. Cañizares, 2013, Electrochemical
324 degradation of the Reactive Red 141 dye using a boron-doped diamond anode: *Water,*
325 *Air, and Soil Pollution*, v. 224.
- 326 Aquino, J. M., M. A. Rodrigo, R. C. Rocha-Filho, C. Sáez, and P. Cañizares, 2012, Influence of the
327 supporting electrolyte on the electrolyses of dyes with conductive-diamond anodes:
328 *Chemical Engineering Journal*, v. 184, p. 221-227.
- 329 Bejan, D., E. Guinea, and N. Bunce, 2012, On the nature of the hydroxyl radicals produced at
330 boron-doped diamond and Ebonex (R) anodes: *Electrochimica Acta*, v. 69, p. 275-281.
- 331 Bergmann, M., A. Kopalal, and T. Iourtchouk, 2014, Electrochemical Advanced Oxidation
332 Processes, Formation of Halogenate and Perhalogenate Species: A Critical Review:
333 *Critical Reviews in Environmental Science and Technology*, v. 44, p. 348-390.
- 334 Bergmann, M., J. Rollin, and T. Iourtchouk, 2009, The occurrence of perchlorate during drinking
335 water electrolysis using BDD anodes: *Electrochimica Acta*, v. 54, p. 2102-2107.
- 336 Carneiro, P., D. Oliveira, G. Umbuzeiro, and M. Zanoni, 2010, Mutagenic activity removal of
337 selected disperse dye by photoelectrocatalytic treatment: *Journal of Applied*
338 *Electrochemistry*, v. 40, p. 485-492.
- 339 Cañizares, P., F. Larrondo, J. Lobato, M. A. Rodrigo, and C. Sáez, 2005a, Electrochemical
340 synthesis of peroxodiphosphate using boron-doped diamond anodes: *Journal of the*
341 *Electrochemical Society*, v. 152, p. D191-D196.
- 342 Cañizares, P., J. Lobato, R. Paz, M. A. Rodrigo, and C. Sáez, 2005b, Electrochemical oxidation of
343 phenolic wastes with boron-doped diamond anodes: *Water Research*, v. 39, p. 2687-
344 2703.
- 345 Cañizares, P., C. Sáez, A. Sánchez-Carretero, and M. A. Rodrigo, 2009, Synthesis of novel
346 oxidants by electrochemical technology: *Journal of Applied Electrochemistry*, v. 39, p.
347 2143-2149.
- 348 Cruz-Gonzalez, K., O. Torres-Lopez, A. Garcia-Leon, E. Brillas, A. Hernandez-Ramirez, and J.
349 Peralta-Hernandez, 2012, Optimization of electro-Fenton/BDD process for
350 decolorization of a model azo dye wastewater by means of response surface
351 methodology: *Desalination*, v. 286, p. 63-68.
- 352 Faouzi, M., P. Cañizares, A. Gadri, J. Lobato, B. Nasr, R. Paz, M. A. Rodrigo, and C. Saez, 2006,
353 Advanced oxidation processes for the treatment of wastes polluted with azoic dyes:
354 *Electrochimica Acta*, v. 52, p. 325-331.
- 355 Guzman-Duque, F., C. Petrier, C. Pulgarin, G. Penuela, and R. Torres-Palma, 2011, Effects of
356 sonochemical parameters and inorganic ions during the sonochemical degradation of
357 crystal violet in water: *Ultrasonics Sonochemistry*, v. 18, p. 440-446.
- 358 He, Z., S. Yang, Y. Ju, and C. Sun, 2009, Microwave photocatalytic degradation of Rhodamine B
359 using TiO₂ supported on activated carbon: Mechanism implication: *Journal of*
360 *Environmental Sciences-China*, v. 21, p. 268-272.
- 361 Hou, M., L. Liao, W. Zhang, X. Tang, H. Wan, and G. Yin, 2011, Degradation of rhodamine B by
362 Fe(0)-based Fenton process with H₂O₂: *Chemosphere*, v. 83, p. 1279-1283.
- 363 Martinez-Huitle, C., E. dos Santos, D. de Araujo, and M. Panizza, 2012, Applicability of diamond
364 electrode/anode to the electrochemical treatment of a real textile effluent: *Journal of*
365 *Electroanalytical Chemistry*, v. 674, p. 103-107.
- 366 Merouani, S., O. Hamdaoui, F. Saoudi, and M. Chiha, 2010, Sonochemical degradation of
367 Rhodamine B in aqueous phase: Effects of additives: *Chemical Engineering Journal*, v.
368 158, p. 550-557.

369 Miller, C., H. Yu, and T. Waite, 2013, Degradation of rhodamine B during visible light
370 photocatalysis employing Ag@AgCl embedded on reduced graphene oxide: *Colloids
371 and Surfaces a-Physicochemical and Engineering Aspects*, v. 435, p. 147-153.
372 Muthukumar, M., D. Sargunamani, and N. Selvakumar, 2005, Statistical analysis of the effect of
373 aromatic, azo and sulphonic acid groups on decolouration of acid dye effluents using
374 advanced oxidation processes: *Dyes and Pigments*, v. 65, p. 151-158.
375 Pandey, A., P. Singh, and L. Iyengar, 2007, Bacterial decolorization and degradation of azo dyes:
376 *International Biodeterioration & Biodegradation*, v. 59, p. 73-84.
377 Panizza, M., and G. Cerisola, 2009a, Direct And Mediated Anodic Oxidation of Organic
378 Pollutants: *Chemical Reviews*, v. 109, p. 6541-6569.
379 Panizza, M., and G. Cerisola, 2009b, Electro-Fenton degradation of synthetic dyes: *Water
380 Research*, v. 43, p. 339-344.
381 Ramirez, C., A. Saldana, B. Hernandez, R. Acero, R. Guerra, S. Garcia-Segura, E. Brillas, and J.
382 Peralta-Hernandez, 2013, Electrochemical oxidation of methyl orange azo dye at pilot
383 flow plant using BDD technology: *Journal of Industrial and Engineering Chemistry*, v. 19,
384 p. 571-579.
385 Rocha, J., A. Solano, N. Fernandes, D. da Silva, J. Peralta-Hernandez, and C. Martinez-Huitle,
386 2012, Electrochemical Degradation of Remazol Red BR and Novacron Blue C-D Dyes
387 Using Diamond Electrode: *Electrocatalysis*, v. 3, p. 1-12.
388 Rodrigo, M. A., P. Cañizares, A. Sánchez-Carretero, and C. Sáez, 2010, Use of conductive-
389 diamond electrochemical oxidation for wastewater treatment: *Catalysis Today*, v. 151,
390 p. 173-177.
391 Saez, C., M. Panizza, M. A. Rodrigo, and G. Cerisola, 2007, Electrochemical incineration of dyes
392 using a boron-doped diamond anode: *Journal of Chemical Technology and
393 Biotechnology*, v. 82, p. 575-581.
394 Serrano, K., P. Michaud, C. Comninellis, and A. Savall, 2002, Electrochemical preparation of
395 peroxodisulfuric acid using boron doped diamond thin film electrodes: *Electrochimica
396 Acta*, v. 48, p. 431-436.
397 Sáez, C., M. A. Rodrigo, and P. Cañizares, 2008, Electrosynthesis of ferrates with diamond
398 anodes: *AIChE Journal*, v. 54, p. 1600-1607.
399 Sánchez-Carretero, A., C. Sáez, P. Cañizares, and M. A. Rodrigo, 2011, Electrochemical
400 production of perchlorates using conductive diamond electrolyses: *Chemical
401 Engineering Journal*, v. 166, p. 710-714.
402 Tavares, M., L. da Silva, A. Solano, J. Tonholo, C. Martinez-Huitle, and C. Zanta, 2012,
403 Electrochemical oxidation of Methyl Red using Ti/Ru_{0.3}Ti_{0.7}O₂ and Ti/Pt anodes:
404 *Chemical Engineering Journal*, v. 204, p. 141-150.
405 Vacca, A., M. Mascia, S. Palmas, and A. Da Pozzo, 2011, Electrochemical treatment of water
406 containing chlorides under non-ideal flow conditions with BDD anodes: *Journal of
407 Applied Electrochemistry*, v. 41, p. 1087-1097.

408

409

410

411

412

413 **Figure Captions**

414 **Figure 1.** Mineralization of Rhodamine B, as a function of Q, during electrolysis in
415 different supporting electrolytes at two current densities (a) 60 mA cm⁻² and (b) 120 mA
416 cm⁻²: ▲ Na₂SO₄; □ HClO₄; ■ H₃PO₄; △ NaCl.

417

418 **Figure 2.** Changes in the COD, as a function of Q, during electrolysis of Rhodamine
419 solutions containing different supporting electrolytes at two current densities (a) 60 mA
420 cm⁻² and (b) 120 mA cm⁻²: ▲ Na₂SO₄; □ HClO₄; ■ H₃PO₄; △ NaCl.

421

422 **Figure 3.** Influence of the current density on the first order kinetic constants for the
423 CDEO of Rhodamine B: ▲ Na₂SO₄; □ HClO₄; ■ H₃PO₄; △ NaCl.

424

425 **Figure 4.** Instantaneous current efficiency constants for the electrolysis of Rhodamine
426 B solutions in different supporting electrolyte media: ▲ Na₂SO₄; □ HClO₄; ■ H₃PO₄;
427 △ NaCl.

428

429 **Figure 5.** Main intermediates detected, as a function of time, during the electrolysis of
430 Rhodamine B solutions containing different supporting electrolytes ((a) HClO₄, (b)
431 NaCl, (c) H₃PO₄ and (d) Na₂SO₄) at 60 mA cm⁻²: phthalic acid (▲); 2,5-
432 hydroxybenzoic acid (●); benzoic acid (■); 3-dinitrobenzoic acid (◆); α- hydroxyglutaric
433 acid (△); intermediate 6 (○) and intermediate 7 (□)

434

435 **Figure 6.** Influence of the current density on the maximum concentration of
436 intermediates detected during the electrolysis of Rhodamine B solutions containing
437 different supporting electrolytes, (a) HClO₄, (b) NaCl, (c) H₃PO₄ and (d) Na₂SO₄:

438 phthalic acid (▲); 2,5-hydroxybenzoic acid (●); benzoic acid (■); 3-dinitrobenzoic acid

439 (◆); α -hydroxyglutaric acid (Δ); intermediate 6 (\circ) and intermediate 7 (\square).

440

441
442
443
444
445
446
447
448
449
450
451
452
453
454
455
456
457
458
459
460
461
462
463
464
465
466

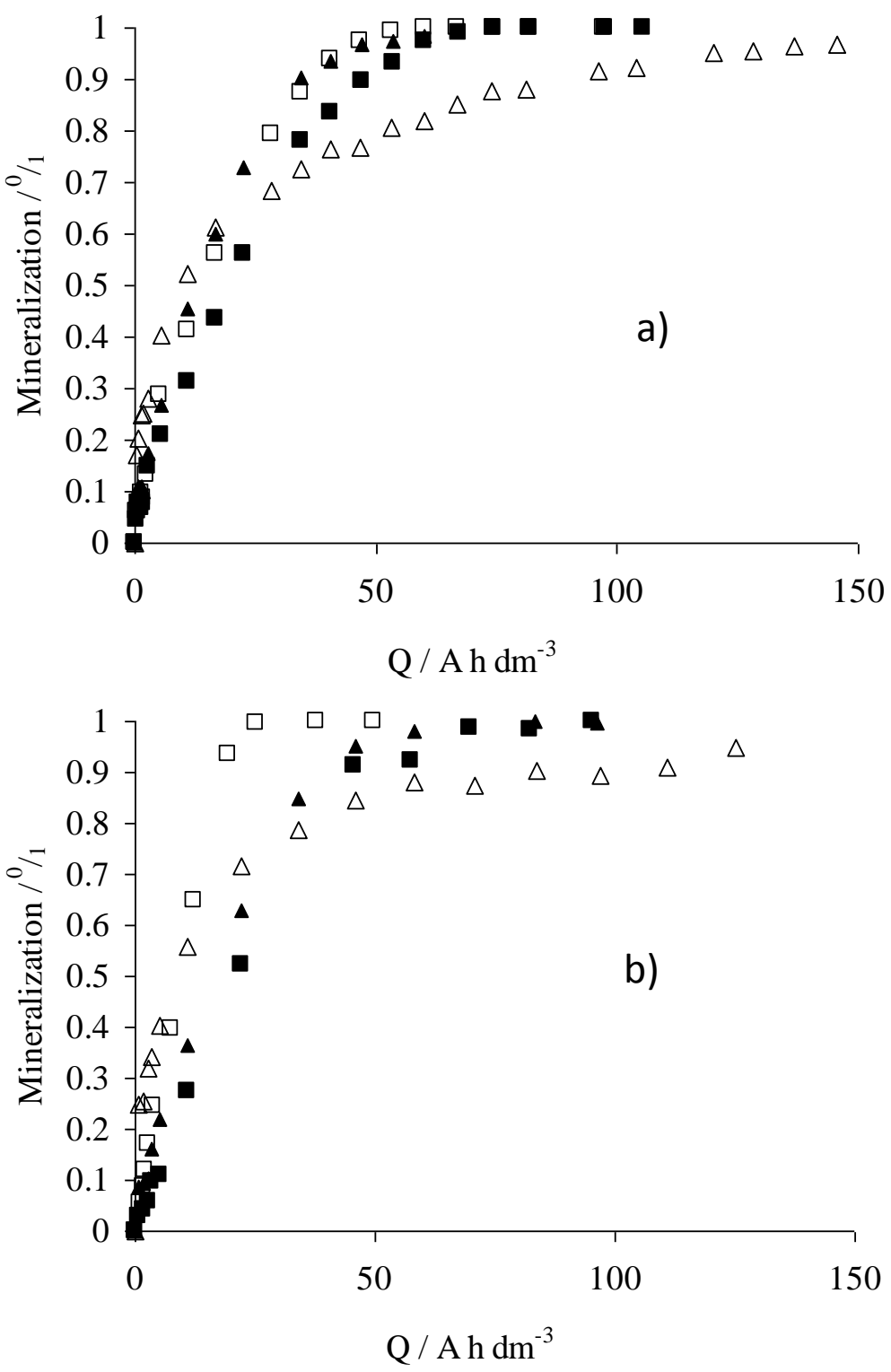


Figure 1

467

468

469

470

471

472

473

474

475

476

477

478

479

480

481

482

483

484

485

486

487

488

489

490

491

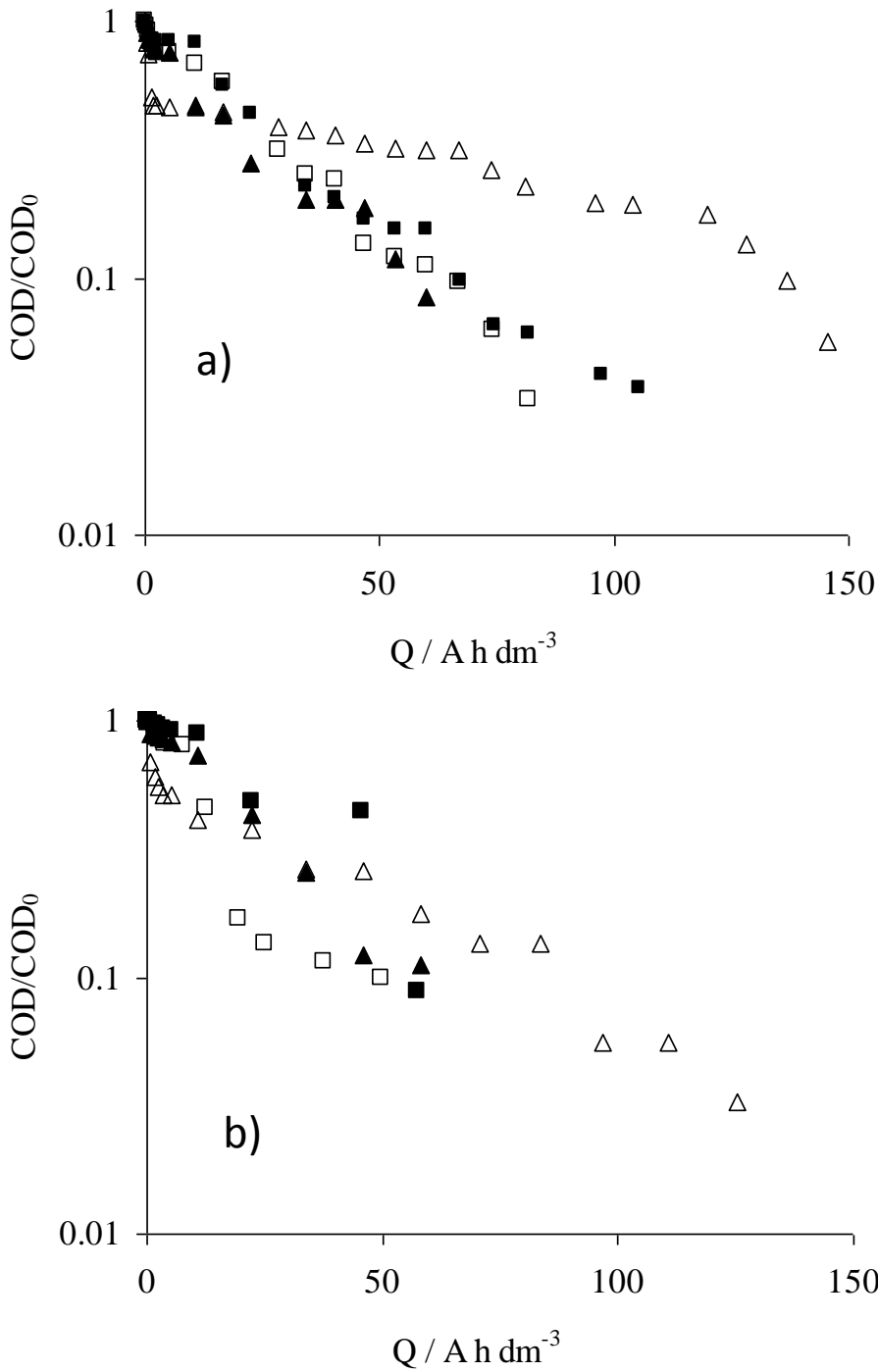


Figure 2

492

493

494

495

496

497

498

499

500

501

502

503

504

505

506

507

508

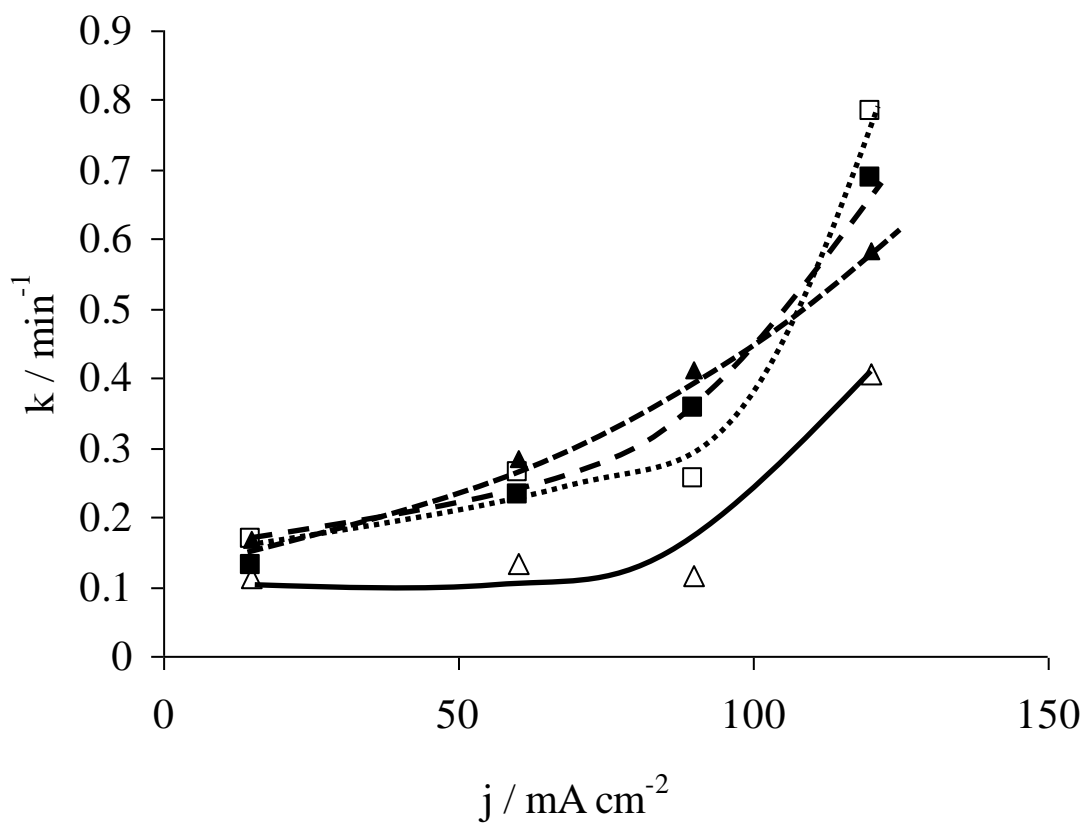


Figure 3

509

510

511

512

513

514

515

516

517

518

519

520

521

522

523

524

525

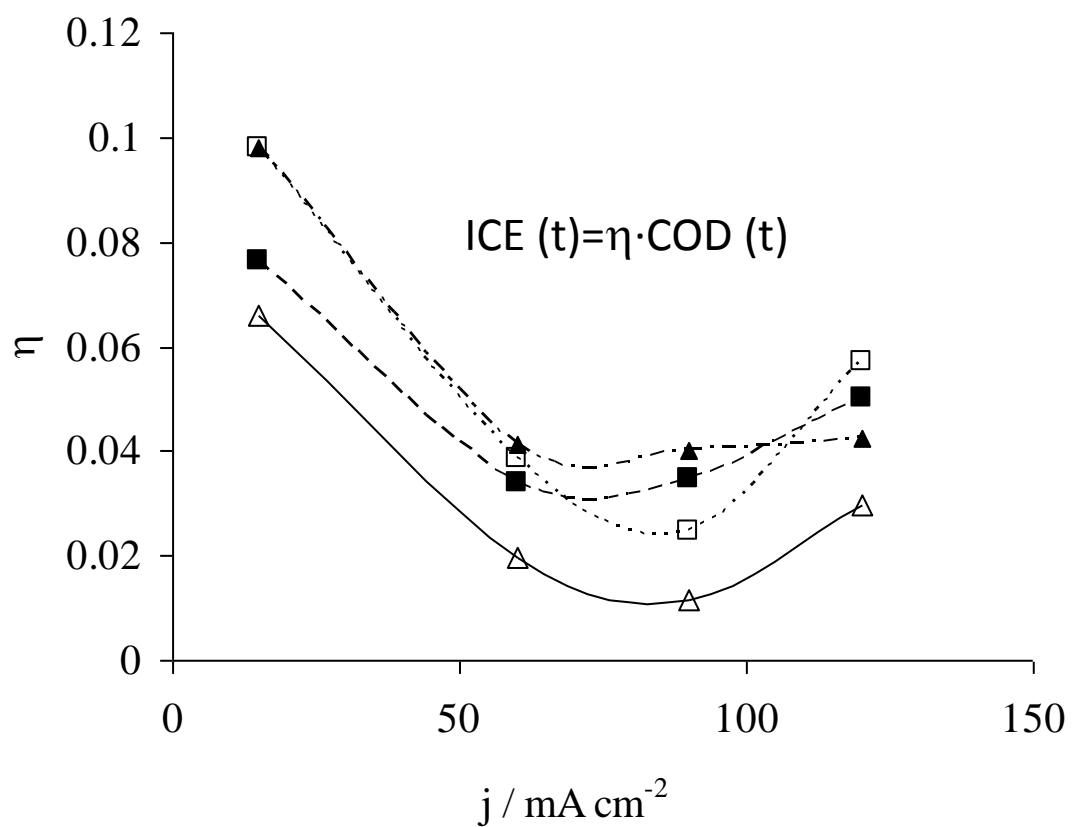
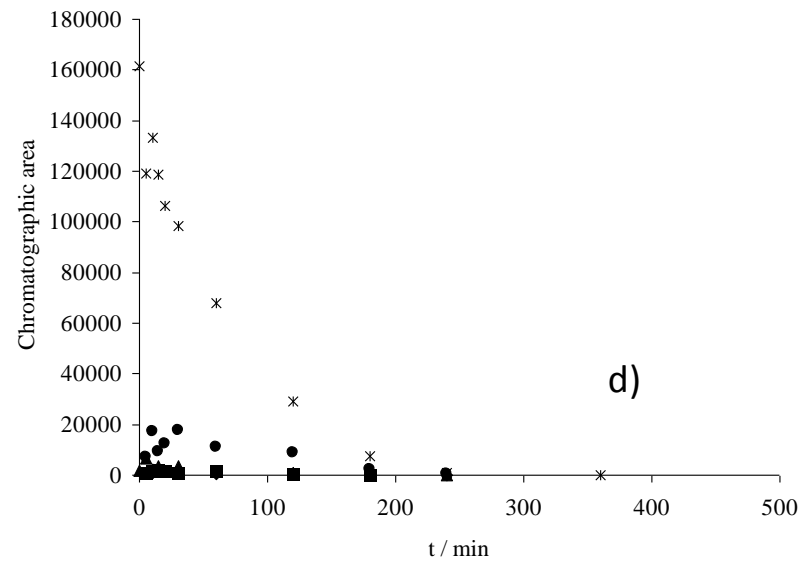
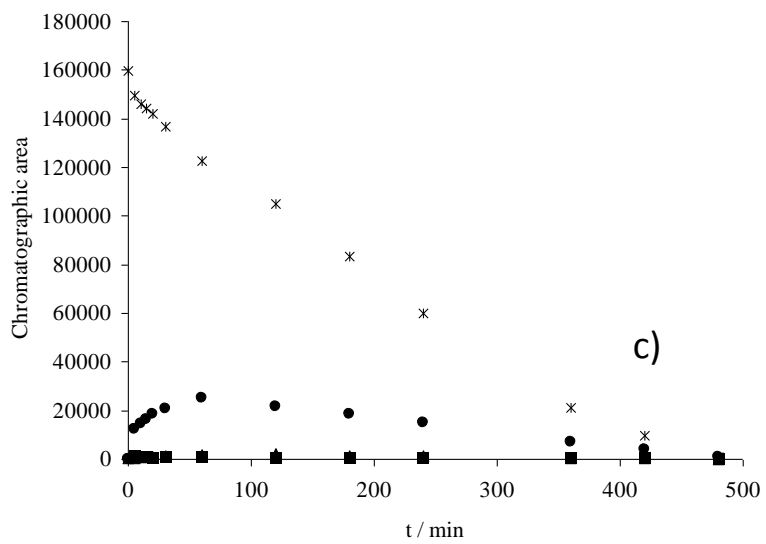
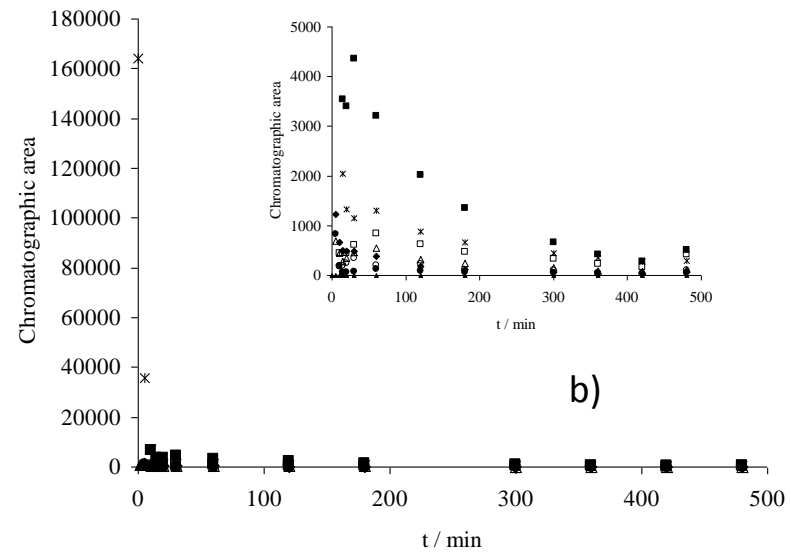
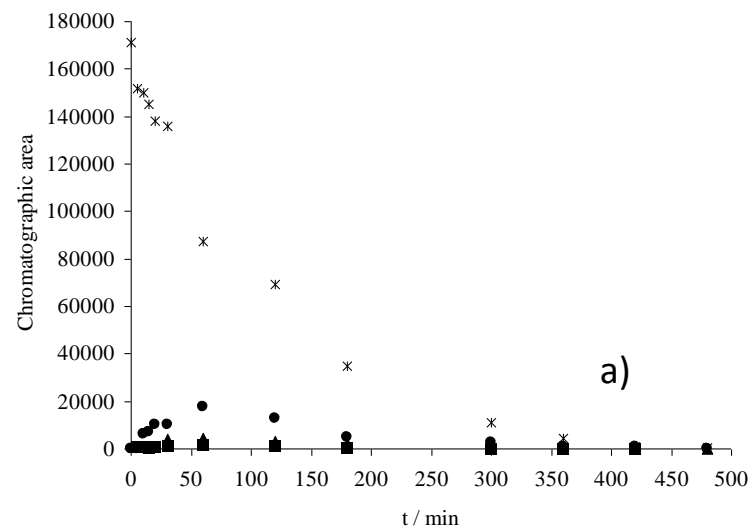
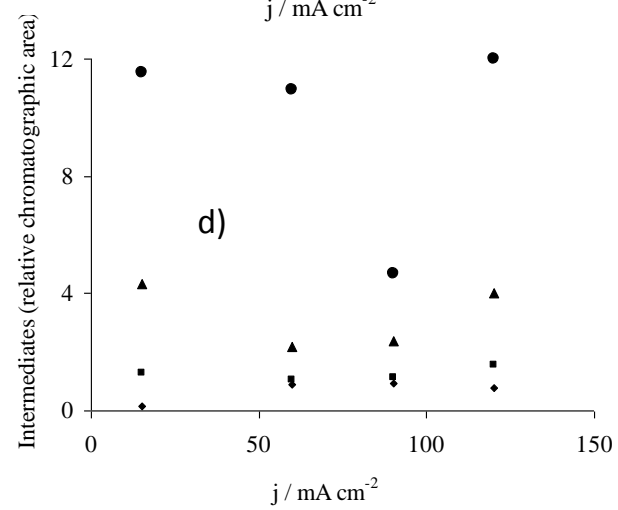
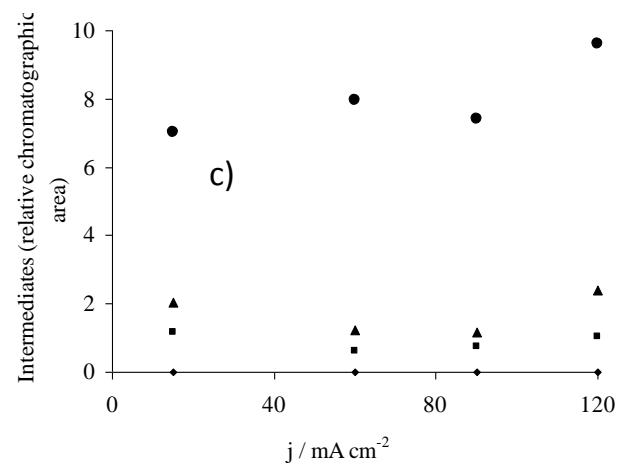
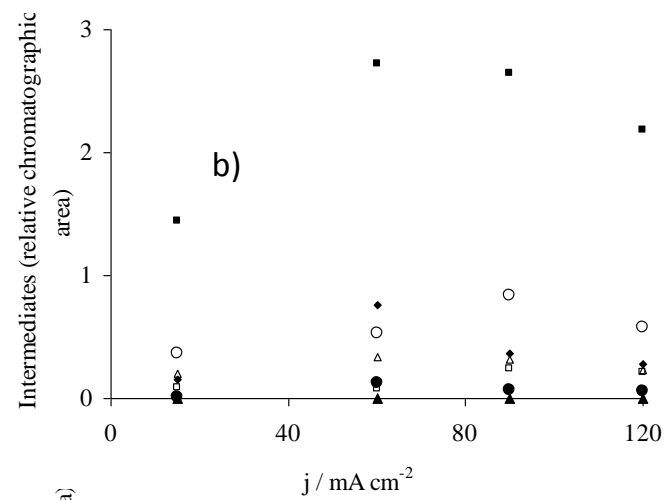
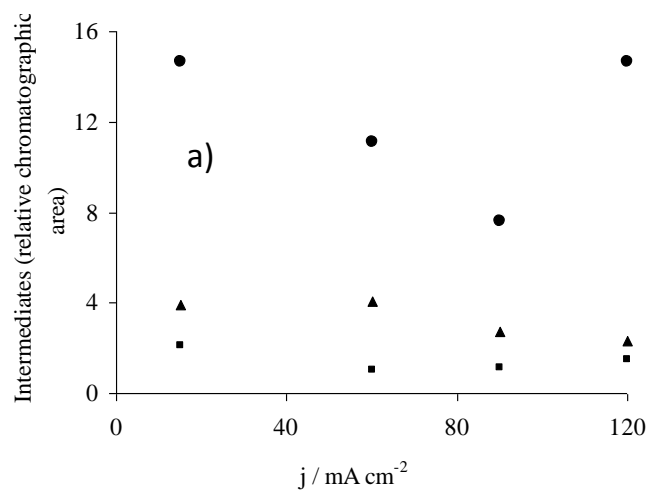


Figure 4





527

528

Figure 6

529

530

531

532

533

534

535

536

537

538

539

540

541

542

543

544

545

546

547

548

549

550

551

552

553

Figure 6

554

555

556

557

558

559

560

561

562

563

564

565

566

567

568

569

570

571

572

573

574

575

576

577

578

579

580

581

582

583

584

585

586

587

588

589

590