

1 **Influence of current density and inlet gas flow in the treatment of gaseous streams**
2 **polluted with benzene by electro-absorption**

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9
10 **Abstract**

11 This paper is focused on the evaluation of the electrochemically assisted absorption process
12 (electro-absorption) as a treatment for volatile organic compounds contained in polluted air
13 flows. Benzene degradation was used to test the technology, using a system consisting of a
14 packed absorption column and a single compartment electrochemical cell equipped with
15 boron-doped diamond (BDD) as anode and steel-steel as cathode. The influence of inlet gas
16 flow rate and operation current intensity was evaluated by monitoring the concentration of
17 benzene and intermediates both, in the liquid absorbent / electrolyte and in the outlet gaseous
18 stream. Results show that steady-state values reached, during benzene absorption and electro-
19 absorption, increase when the inlet gas flow rises from 1.5 to 6.0 L h⁻¹ and that the bare
20 absorption process retains benzene only up to this steady-state value of the liquid and then,
21 it has no effect on the removal of benzene. Concentrations of benzene in the
22 absorbent/electrolyte decreases with the increase in the current density during
23 electrochemically assisted processes. As well, mineralization was found to be very important,
24 as expected because of the very low concentrations of benzene into the liquid and the very

25 high efficiencies of the electrolytic degradation with BDD electrodes, and benzoquinone and
26 carboxylic acids were identified as the primary intermediates. These species were not
27 detected in the outlet gas flow. The decrease in the concentration and mass flow of benzene
28 in the outlet streams confirms that electro-absorption is a functional electrochemical
29 application to remove benzene from gaseous streams. When current density rises (10, 50 and
30 100 mA cm⁻²), the degradation percentage increases. The values obtained were 43, 61 and
31 75 % with 1,5 L h⁻¹ of inlet gas flow and 10, 25 and 37 % with 6 L h⁻¹. However, the
32 degradation percentage relative to the energy consumption decreases when the current
33 density rises. Results were, correspondingly, 24.7, 4.2 and 2.2 % kWh⁻¹ with 1,5 L h⁻¹ of
34 inlet gas flow and 7.1, 2.1 and 1.4 % with 6 L h⁻¹. This energy efficiency trend was explained
35 because mass transport becomes the bottleneck in the oxidation of diluted organic solutions.
36 These results contribute to expand the understanding of the mechanisms involved in the
37 electro-absorption of benzene and recommend for further research in the topic, due to the
38 relevance of the treatment results.

39

40 **Keywords**

41 Electro-absorption, benzene, boron-doped diamond, current density, inlet gas flow.

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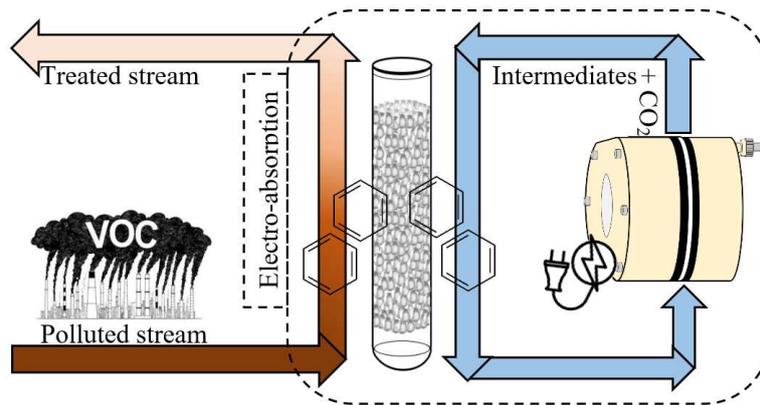
43 **Highlights**

- 44 • Efficient mineralization of the VOCs contained in the air using electro-scrubbing
45 technology.
- 46 • Benzoquinone and carboxylic acids are formed in the absorbent but not found in the outlet
47 stream.

- 48 • Percentage of benzene gaseous removal increases by increasing operation current density.
- 49 • Mass transport control of the electrochemical degradation of benzene into the
- 50 absorbent/electrolyte.

51

52 **Graphical Abstract**



53

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56

57 **1. Introduction**

58 In the last decades, volatile organic compounds (VOCs) had constituted a relevant group of
59 chemicals used in different processes in several industries, such as paints and solvents [1],
60 rubbers [2], wooden furniture [3], anode manufacturing [4], polyester fabric [5] and asphalt
61 pavement [6], among many others. These activities are associated with an environmental
62 problem that is becoming of major relevance: the generation of gaseous polluted streams [7]
63 which influence not only human health but also air quality and tropospheric chemistry [8].
64 Nowadays, VOCs contained in polluted air are mainly destroyed by different oxidation
65 processes including thermal, catalytic, and biological [9–11].

66 In recent years, very efficient electrochemical applications have been developed to treat
67 liquid wastes and it has been pointed out the high impact of the choice of the anodic material
68 [12–15] and operation conditions on the performance attained [16–20]. These interesting
69 results have pushed researchers to expand the scope to the treatment of gaseous pollutants.
70 An innovative technology is electro-absorption (also called electro-scrubbing). It combines
71 physical or chemical absorption and electro-oxidation. The first step allows the molecules of
72 the pollutant to transport from the gaseous to the bulk of the liquid phase. Principal
73 parameters to consider are the type of packing, column dimensions, transfer coefficients and
74 feed flowrates [21]. The second step is carried out in an electrochemical cell and enables the
75 elimination of the pollutants by the oxidation reactions which can take place on the surface
76 of the electrodes and/or in the liquid bulk by the oxidants produced. Important parameters in
77 electro-oxidation are electrodes material, electrolyte, current density and cell voltage [22–
78 25]. This process is now developed at technology readiness levels (TRL) of 3-4, and the
79 interest has been mainly focused on the removal of VOCs and odour substances [26–30] but
80 applicability may be extended also to atmospheric pollutants. Hence, there is a great need for

81 further research up to meet an industrial TRL of 7-9, and in the incoming years many
82 scientific and technically sound works are expected on this topic. Some advantages of
83 electro-absorption as compared to in-use technologies are the high efficiency of anodic
84 oxidation, the regeneration of the absorbent solution, the possibility of optimizing the
85 formulation of the absorbent and the fact that the primary reagents used are the electrons.
86 Considering this background, this study aims to evaluate the electro-scrubbing of streams of
87 air polluted with benzene, where this species has been used as a model of VOC. Packed
88 absorption column and single compartment electrochemical cell are combined using anode
89 consisting of boron-doped diamond and the effect of the current densities for two different
90 feeding flowrates has been clarified, paying attention to the intermediates formed and the
91 decrease of benzene concentration in the polluted gas stream after the application of the
92 electro-absorption process.

93

94 **2. Experimental**

95 **2.1 Chemicals**

96 All the chemical reagents were used as received and were analytical grade. The main ones
97 were benzene 99.9% (Sigma-Aldrich), sulfuric acid 97% and hexane HPLC grade (Scharlau).
98 Also, the following potential intermediates and by-products typically reported in the
99 oxidation of benzene [31–34] were used: benzoquinone 98% (Panreac), phenol 99% (Sigma-
100 Aldrich), maleic acid 99% (Panreac), malonic acid 99% (Sigma-Aldrich), acetic acid 96%
101 (Panreac), formic acid 98% (Sigma-Aldrich) and oxalic acid 99% (Sigma-Aldrich). All
102 solutions were prepared with deionized water (Millipore Mili-Q system, resistivity 18.2 M Ω
103 cm at 25°C). Helium (Al Air Liquide España), filtered by a hydrocarbon cartridge filter

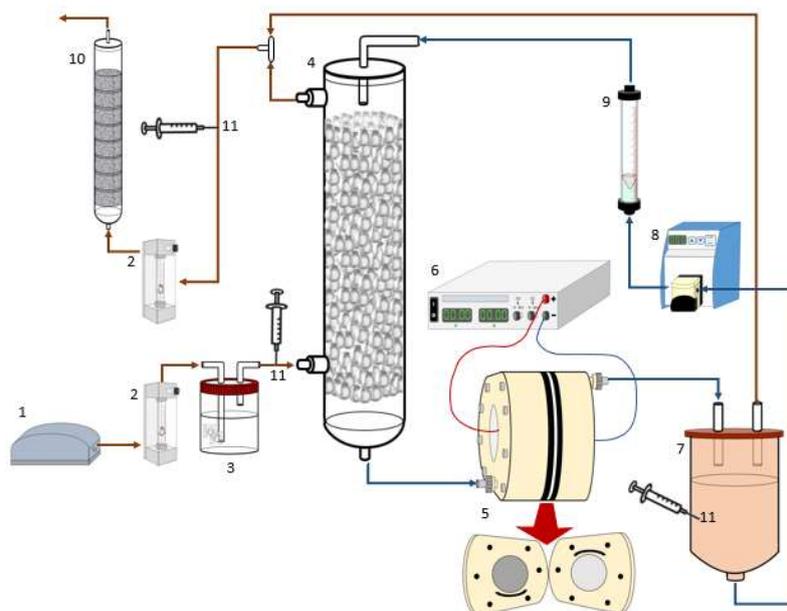
104 (Thermo Fisher Scientific), was used for the gas chromatography. Active carbon was BD
105 CARBON GR 10.3 (Barcelonesa).

106

107 **2.2 Experimental setup and procedure**

108 A commercial electrochemical cell (DiaCell, purchased to Adamant Technologies, France)
109 and a packed absorption column were integrated to constitute the electrochemically assisted
110 absorption system shown in Figure 1 (electro-absorber). The column was made of glass with
111 an inner diameter of 0.05 m, length of 0.5 m, and a packing height of 0.4 m, and it was filled
112 with glass spheres (8 mm diameter) randomly distributed. A synthetic polluted gaseous
113 stream was obtained by bubbling air (1) [35] into a tank filled with liquid benzene (99,9%)
114 (3). The gas flowrate was monitored with a rotameter (2) and the inlet of the gas was placed
115 in the bottom of the absorption column (4). The absorbent liquid was contained in a tank
116 (1L, 0.05 M H₂SO₄), from which it was pumped to the top of the column by a peristaltic
117 pump (8) with a flow (10 L h⁻¹) monitored with a rotameter (9). Both streams had
118 countercurrent contact (7). The liquid stream that comes out from the bottom of the column
119 went to a single-compartment electrochemical flow cell (5). The anode is boron-doped
120 diamond (BDD) with Si as support and the cathode is commercial stainless steel (SS), both
121 having a circular shape with an area of 78 cm². The distance between the electrodes is 5 mm.
122 A power supply (6) provides a constant intensity to the cell during each of the tests carried
123 out. The liquid stream that comes from the cell was recirculated to the reactor tank, while its
124 gaseous stream and the outlet gas from the column were mixed and conducted to an active
125 carbon column to retain unreacted benzene and intermediates from the system. In this way,
126 it can be considered that the gaseous phase performs in continuous mode (red line) while the
127 liquid phase works in batch mode (blue line).

128 The flowrates of gas treated evaluated were 1.5 and 6.0 L h⁻¹ and current densities applied
129 were 10, 50 and 100 mA cm⁻². Absorption corresponds to experimentation with 0 mA cm⁻².
130 Each experimental test was carried out with a new electrolyte solution.
131



132
133 Figure 1. Electro-absorber installation scheme. (1) air compressor, (2) rotameter, (3) liquid
134 benzene tank, (4) packed absorption column, (5) flow electrochemical cell with BDD as
135 anode and stainless steel as cathode, (6) power supply, (7) reactor tank, (8) peristaltic pump,
136 (9) flow meter, (10) active carbon column, (11) sampling points.

137

138 2.3 Analytical techniques

139 From the reactor tank, liquid samples were collected (2.0 mL). They underwent L-L
140 extraction with hexane (6 mL) in a vortex stirrer for 2 minutes. After the two phases were
141 split, hexane was transferred to a gas chromatography (GC) vial. From the absorption column
142 and the activated carbon column inlets, gaseous samples (5.0 mL) were taken and bubbled
143 directly in hexane (10.0 mL). Then, they were collected in a GC vial. Compound

144 concentrations in the samples were measured by Shimadzu Gas Chromatography (Nexis GC-
145 2030) coupled to a Mass Spectrometer Detector (GCMS-QP2020 NX). The capillary
146 analytical column was a SH-Rxi-5Ms 30m x 0.25 mm x 0.25um with a split/splitless injector
147 in split mode (ratio of 20) and a total flow of 42.8 mL min⁻¹ (pressure 100.0 kPa) with a linear
148 velocity of 48.3 cm s⁻¹. Temperature column program was: 38.0 °C for 3 min, 40.0 °C min⁻¹
149 to 75.0 °C and 30.0 °C min⁻¹ to 200.0 °C for 2 min. The injection and detector temperatures
150 were 200.0 °C. The carboxylic acids were measured by HPLC with a Hi-Plex H column
151 (Agilent Technologies) and a 5.0 mM sulfuric acid aqueous solution as the mobile phase (0.8
152 mL min⁻¹). The wavelength was 210 nm. Total organic carbon (TOC) was measured by Multi
153 N/C 3100 analyser (Analytik Jena).

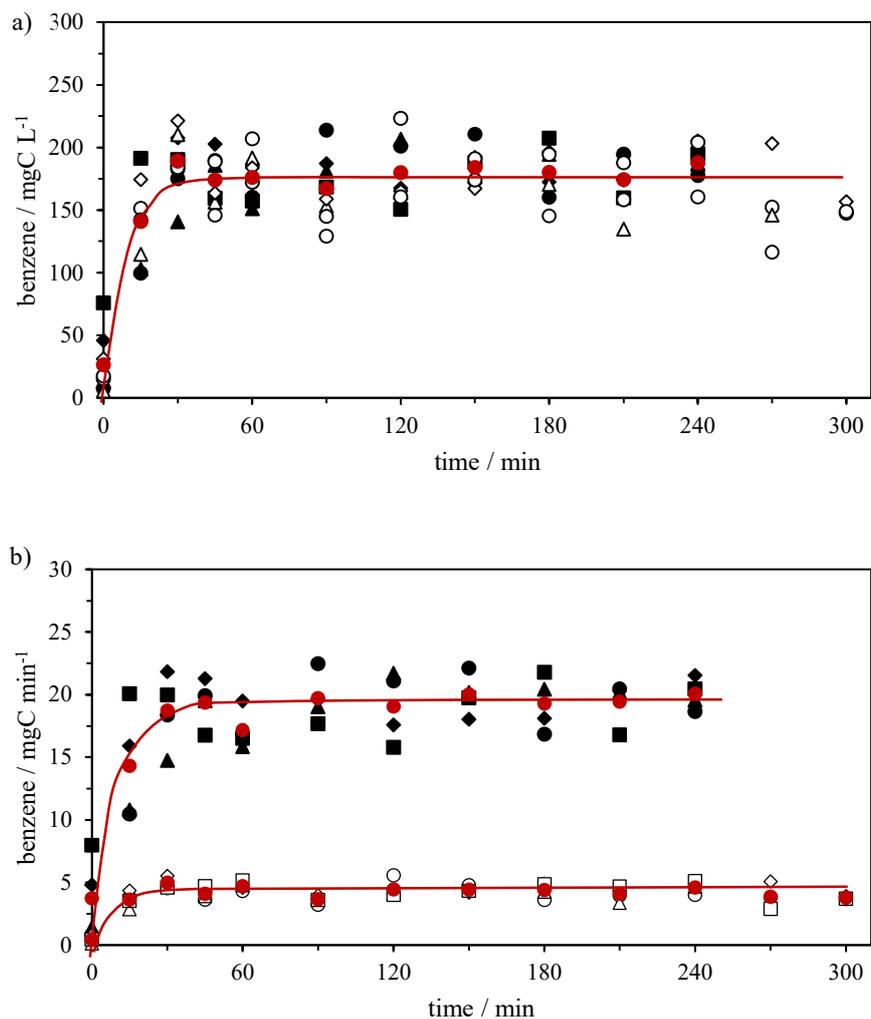
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155 **3. Results and Discussion**

156

157 Figure 2 shows the concentration and mass flow of benzene fed into the absorption (0 mA
158 cm⁻²) and electro-absorption processes, in which the effect of the carrier flowrate (1.5 and
159 6.0 L h⁻¹) and operating current densities (10, 50 and 100 mA cm⁻²) were evaluated.
160 Obviously, the inlet is not influenced by the treatment applied, although it gives relevant
161 information about the variability of the inputs in this experimental device. Thus, as it can be
162 observed, there is a fluctuation of nearly 15% in the inlet measurements (188 ± 28 mg C L⁻¹
163 ¹), that can be explained in terms of the high turbulence produced by the bubbling of air in
164 the benzene tank. This dispersion is also observed in the mass flow fed and the average values
165 were 4.23 ± 0.63 and 19.21 ± 2.02 mg C min⁻¹ at 1.5 and 6 L h⁻¹ respectively.

166



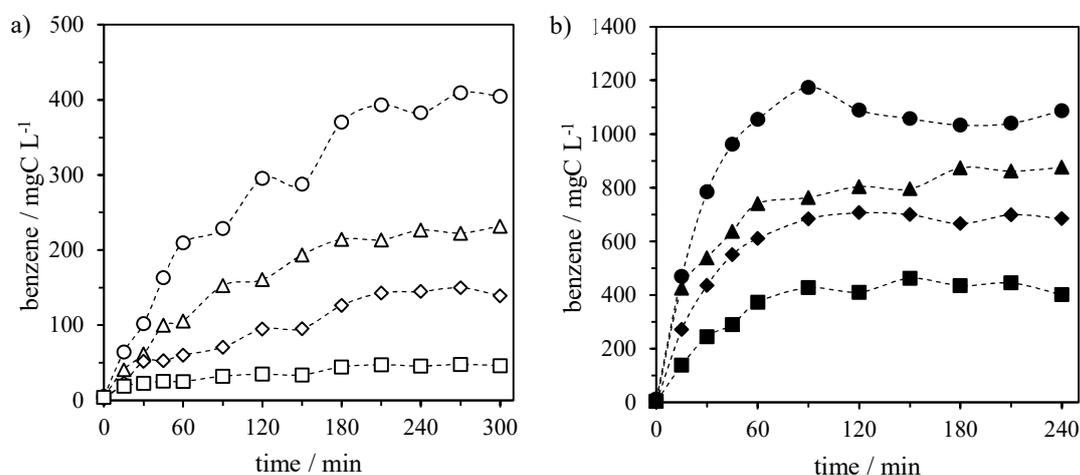
167 Figure 2. Time evolution of benzene a) concentration and b) mass flow in inlet gaseous flow
 168 of 1.5 (white) and 6 (black) L h^{-1} . Averages of every time (red points) and their trend line
 169 (red line). Absorption (circle) and electro-absorption processes at 10 (triangle), 50 (rhombus)
 170 and 100 (square) mA cm^{-2} .

171

172 Figure 3 shows the time course of the benzene concentration in the absorbent liquid phase
 173 during the absorption and electro-absorption tests. Regarding the non-electrochemical
 174 process, as expected, only benzene is measured in the liquid and no reactivity is associated.
 175 This is because of the low reactivity of VOCs under absorption conditions. A similar behavior

176 was observed in the degradation of toluene [26]. Physical adsorption of benzene in water is
177 not affected by the concentration of the electrolyte used (0.05 M H₂SO₄) [36]. It mainly
178 depends on equilibrium (and hence on the temperature and pressure) and mass transport
179 (where flowrates are expected to influence). Thus, since temperature and pressure were kept
180 constant during the tests (25°C and 93 kPa), the latest justifies that the amount of benzene
181 absorbed by the system grows by increasing the gas inlet flow (mass flow). Steady-state value
182 was around three times lower in the absorbent of the system fed with 1.5 than with 6.0 L h⁻¹
183 of benzene polluted air, being the values in which the concentration stabilizes of 392.12 ±
184 16.01 and 1080.04 ± 51.48 mg C L⁻¹, respectively. Additionally, it can be observed that the
185 concentration in the liquid reached a plateau after about three hours of feeding with 1.5 L h⁻¹
186 of benzene polluted air and only one hour and a half of feeding the system with 6.0 L h⁻¹.
187 More interesting are the stabilization concentrations reached in the electro-absorption, that
188 is, when current is applied in the electrochemical cell connected to the absorption column:
189 the benzene concentration achieved at steady-state in the liquid phase decreases with the
190 increase in the current density. With a flow of 1.5 L h⁻¹, the values are 221.90 ± 7.79, 140.75
191 ± 8.85 and 46.40 ± 1.30 mg C L⁻¹ in the electro-absorptions at current densities of 10, 50 and
192 100 mA cm⁻², respectively. In the tests in which the treatment technology is fed with 6.0 L
193 h⁻¹, the concentrations reached with the same current density values were 829.06 ± 47.42,
194 690.02 ± 14.64 and 430.23 ± 22.38 mg C L⁻¹. When these values are compared with the
195 obtained in the bare absorption, it is observed higher percentages of reduction in the lowest
196 inlet gas flow tested. Thus, reduction values were 43.41, 64.11 and 88.17 % with 10, 50 and
197 100 mA cm⁻². On the other hand, the tests with the highest inlet gas flowrate achieved a
198 23.24, 36.11 and 60.17 % of reduction in the stabilization concentration for the three
199 increasing current densities tested in this work. This could happen because while the current

200 density increments, the production rate of oxidants increases as well. Consequently, as there
 201 is a greater amount of these reactants, the kinetics of benzene degradation in the liquid phase
 202 is higher and its steady-state value is lower. Additionally, it can be concluded that with an
 203 increase of four times in the inlet gas flow rate, the percentage of reduction in the steady-
 204 state concentration in the liquid phase is reduced from 1.5 to 1.9 times for the same value of
 205 current density.
 206

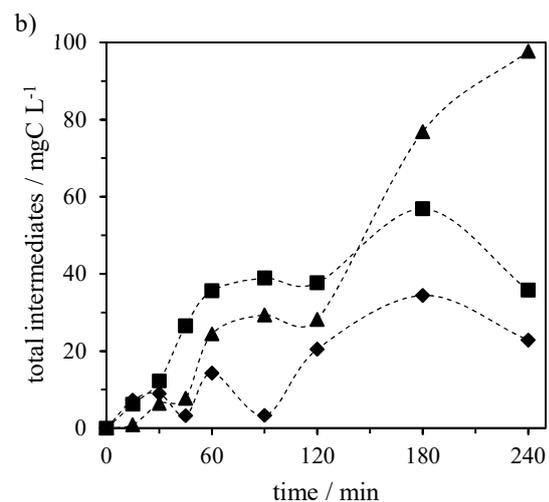
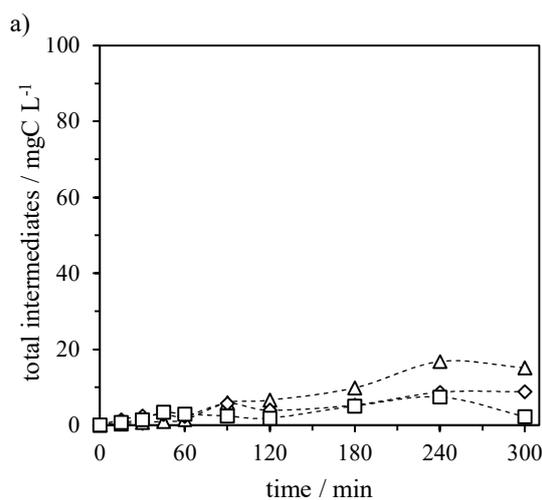


207 Figure 3. Time evolution of benzene concentration in liquid phase in absorption (circle) and
 208 electro-absorption processes at 10 (triangle), 50 (rhombus) and 100 (square) mA cm⁻². At
 209 different inlet gas flows of a) 1.5 (white) and b) 6 (black) L h⁻¹.

210

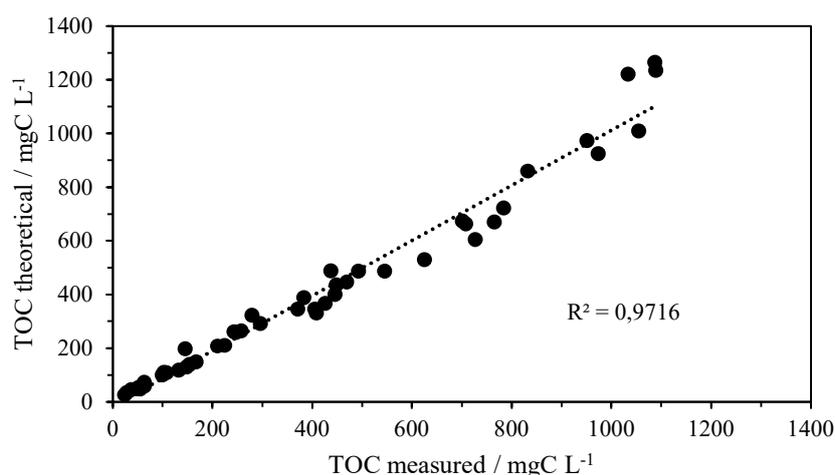
211 Figure 4 shows the time course of the total concentration of the intermediate compounds
 212 identified analytically in the liquid phase in electro-absorption processes at inlet gas flows of
 213 1.5 and 6.0 L h⁻¹. The time evolution of each compound is in Figure S1. In both flow rates,
 214 benzoquinone was the principal aromatic intermediate identified, which reached maximum
 215 concentrations of 8.0 and 44.0 mg C L⁻¹ for each inlet gas flow fed, respectively.

216 Additionally, carboxylic acids identified were oxalic, maleic, malonic and acetic. The
217 concentration of the first one is higher when the current density increases, and the same
218 behaviour happens with malonic acid. On the other hand, the other acids show the highest
219 concentration at 10 mA cm⁻², while at 50 and 100 mA cm⁻² their concentrations are much
220 smaller. In the tests with 1.5 L h⁻¹ of inlet gas flow, all maximum concentrations of carboxylic
221 acids are lower than 10 mg C L⁻¹, while with 6 L h⁻¹ they do not exceed 40 mg C L⁻¹. The
222 carbon mass of the intermediates represents less than 10% of the total mass that enters in the
223 system in both inlet gas flows. So, this fact could infer that there is not an accumulation of
224 intermediates and mineralization is the principal way of the benzene electro-degradation
225 process. Also, these results help to have information about the degradation route of benzene
226 which could consist first in the oxidation of the aromatic ring before its breakup and
227 subsequent formation of carboxylic acids to finally reach mineralization. This statement is in
228 concordance with other authors [37–40].
229



230 Figure 4. Time evolution of the total concentration of intermediate compounds identified in the liquid
231 phase at a gaseous flow of a) 1.5 (white) and b) 6 (black) L h⁻¹ in electro-absorption processes at 10
232 (triangle), 50 (rhombus) and 100 (square) mA cm⁻²
233

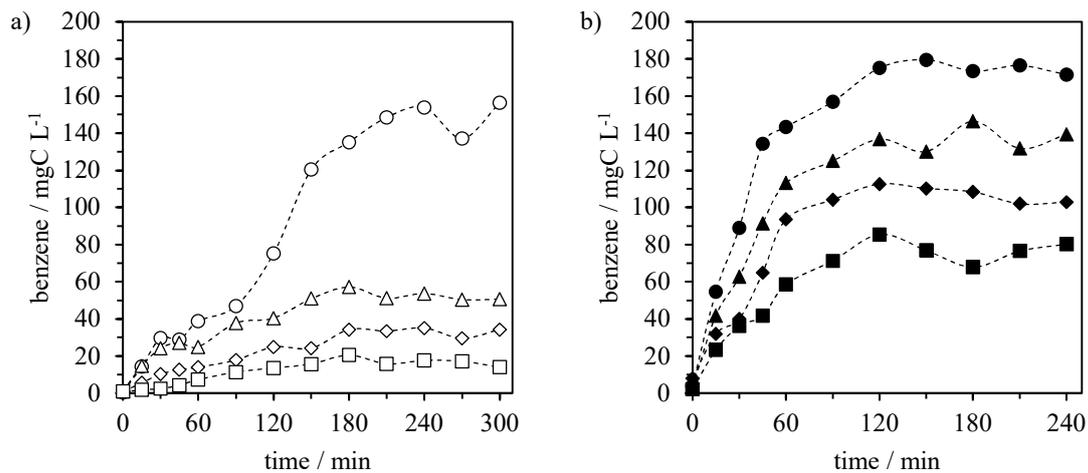
234 Total organic carbon (TOC) was calculated by the sum of benzene and intermediates
235 compounds concentrations, and it is compared with TOC measured analytically, results are
236 shown in Figure 5. It is observed that theoretical and measured TOC values have a good
237 fitting coefficient (R^2 0.97). This means that the mass of carbon missed from the input value
238 is mineralized and that there are no other relevant intermediates unidentified that are
239 interfering with TOC measurement.



240
241 Figure 5. TOC theoretical vs measured in liquid phase in absorption and electro-absorption
242 processes at 10, 50 and 100 mA cm⁻² at different inlet gas flow of 1.5 and 6.0 L h⁻¹.
243

244 Values shown in figures 3 to 5 are important to understand the performance of the electro-
245 absorption technology. However, the most important information comes from the
246 comparison of the concentrations and mass flow of benzene in the gaseous outlet streams.

247 Figure 6 shows the concentration of benzenes detected in this stream during the different
 248 tests and Figure 7 collects the mass flow of outlet gas in absorption and electro-absorption
 249 tests at gas flow rates of 1.5 and 6.0 L h⁻¹. The first Figure confirms that concentrations of
 250 benzene in the treated gas reach a steady-state value in less than three hours, which
 251 corresponds to the same concentration of the inlet (Figure 2) in the case of the bare absorption
 252 technology (as expected, because of the steady-state value reached in the absorbent capacity)
 253 and too much lower values in the concentration of benzene in the outlet gas of the electro-
 254 absorbers when the electric current is applied, as explained because of the electrochemical
 255 destruction of benzene and formation of intermediates.
 256



257 Figure 6. Time evolution of benzene concentration in outlet gas flow in absorption (circle)
 258 and electro-absorption processes at 10 (triangle), 50 (rhombus) and 100 (square) mA cm⁻².
 259 At different inlet gas flows of a) 1.5 (white) and b) 6 (black) L h⁻¹.

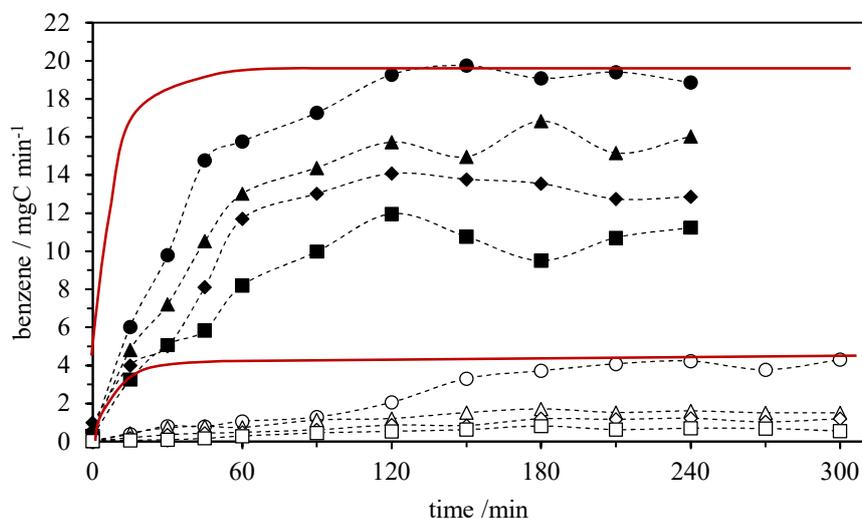
260

261 The decrease in benzene concentrations in the outlet gas is also reflected in the observed
 262 differences in the mass flow rate escaping treatment (Figure 7). Thus, it is observed that when
 263 absorption is carried out, the benzene concentration gradually increases until the input value

264 is reached. This fact means that liquid is no longer able to absorb benzene and, this
265 phenomenon happens about three hours with the 1.5 L h⁻¹ flow rate and one hour and a half
266 with 6.0 L h⁻¹. It could be explained because the equilibrium concentration between liquid
267 and gaseous phases strongly depends on the partial pressure, mass flow in other words, of
268 the target compound within the gas phase in absorption. In addition, increasing the gas flow
269 induces more turbulence inside the column that leads to better interface contact between the
270 liquid and gas phases improving the mass transfer rate. However, the higher the gas flow, the
271 shorter is the residence time of the gas in the system, which affects the diffusivity of the
272 pollutant [41]. So, it is important to find an optimal point between these parameters. The
273 steady-state flow was 4.02 ± 0.26 and 18.94 ± 0.87 mg C min⁻¹ and the difference in
274 percentage in relation with the inlet mass flow was 2.20 and 3.56 % respectively with 1.5 and
275 6.0 L h⁻¹ of inlet gas flow rate.

276

277 Regarding the electro-absorption tests, it is clearly observed that the higher is the current
278 density, the lower is the mass flow value that the outlet gas reaches in the steady-state. With
279 a flowrate of polluted air of 1.5 L h⁻¹ and current densities of 10, 50 and 100 mA cm⁻², the
280 outlet gas mass flows were 1.58 ± 0.09 , 1.17 ± 0.08 and 0.68 ± 0.10 mg C min⁻¹ respectively.
281 Meanwhile, with 6.0 L h⁻¹, these values are 15.52 ± 0.87 , 13.34 ± 0.54 and 10.70 ± 0.88 mg
282 C min⁻¹. When these results are compared with the inlet mass flow in each test at different
283 current densities, the reduction percentages are 61.53, 73.64 and 84.09 % with the lowest
284 inlet gas flow and 18.48, 32.50 and 41.86 % with the highest one. In a similar way to the
285 liquid phase, it can be determined that with an increase of four times in the inlet gas flow
286 rate, the percentage of reduction in the steady-state concentration in the outlet gas mass flow
287 is reduced from 2.0 to 3.3 times for the same value of current density.



289

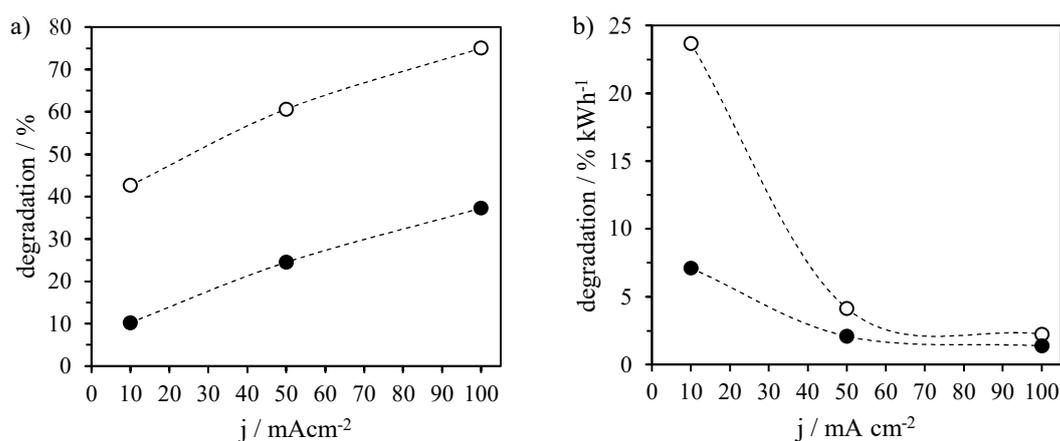
290 Figure 7. Time evolution of benzene mass flow in outlet gas stream in absorption (circle) and
 291 electro-absorption processes at 10 (triangle), 50 (rhombus) and 100 (square) mA cm^{-2} . At
 292 different inlet gas flows of a) 1.5 (white) and b) 6.0 (black) L h^{-1} . Mass flow of benzene in
 293 the inlet gas (red line in Figure 2b).

294

295 The mass of benzene adsorbed in the active carbon column which represents the amount not
 296 treated is shown in Table S1. In addition, no intermediates were identified in the gases
 297 outflowing from the system, which can be explained by the very low concentrations
 298 measured in the absorbent liquids and, also, by the very low volatility of all the intermediates.
 299 This supports the technical feasibility of the treatment proposed for these gases containing
 300 VOCs as has been demonstrated with other compounds [24,42–44]. As well, energy and
 301 efficiency considerations are important to be discussed and this information is provided in
 302 Figure 8, where Part a indicates that the degradation percentage rises with current density
 303 and Part b, which represents the degradation percentage relative to kWh (calculated with the

304 voltage, intensity and total time of each test) that energy efficiency decreases very
305 importantly with current density. This lower efficiency at higher current densities is expected
306 because of the mass transport control of the electrochemical process in liquid phase which,
307 in turn, can be associated to the low concentration of pollutants.

308



309 Figure 8. Degradation percentage a) absolute and b) relative to kWh at different current
310 densities. Inlet gaseous flow of 1.5 (white) and 6.0 (black) L h⁻¹.

311

312 CONCLUSIONS

313 Conclusions obtained in this study are the following:

- 314 • Electrochemically assisted absorption processes are efficient in the degradation of
315 benzene contained in polluted air and may become a very interesting choice if the
316 TRL increases. Positive results obtained in this work are promising a recommend
317 further work in the topic.
- 318 • Polluted air can be simulated by bubbling air into liquid benzene. Inlet gas
319 concentration is constant and independent of the gas flow rate, in contrast to the mass
320 flow which is proportional with this parameter. Additionally, steady-state

321 concentration of benzene absorbed increases and is reached in half the time when
322 inlet gas flow rises from 1.5 to 6.0 L h⁻¹.

323 • Intermediates compounds identified during electro-absorption processes are
324 benzoquinones and carboxylic acids such as oxalic, maleic, malonic and acetic. Their
325 concentration is very low, as expected for an electrochemical oxidation of organics
326 with boron-doped diamond electrodes. Therefore, mineralization is considered the
327 primary degradation mechanisms of benzene in electro-absorption process.

328 • Electro-absorption process is functional to the treatment of gaseous streams polluted
329 with benzene, a higher percentage of benzene removal from the gaseous stream is
330 obtained by increasing the current density and decreasing the mass flow of inlet gas
331 flow. The same behaviour is observed in the liquid phase.

332 • Degradation percentage relative to energy consumption increases when the current
333 density decrease, this energy efficiency relation with current density is due to the
334 limitations in mass transport control during the oxidation of diluted aqueous wastes.

335

336 **Author contributions**

337 **Andrea N. Arias:** Investigation, Data curation, Formal analysis, Writing – original draft. **R.**

338 **Granados-Fernández:** Writing – review & editing. **J. Lobato:** Funding acquisition, Project
339 administration, Supervision, Validation, Writing – review & editing. **Carmen M.**

340 **Fernández-Marchante:** Supervision, Validation, Writing – review & editing. **Manuel A.**

341 **Rodrigo:** Conceptualization, Funding acquisition, Project administration, Supervision,
342 Validation, Writing – review & editing.

343

344 **Declaration of competing interest**

345 The authors declare that they have no known competing financial interests or personal
346 relationships that could influence the work reported in this paper.

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351

352 **References**

- 353 [1] Z. Liu, Y. Yan, T. Lv, Z. Huang, T. Liu, Q. Huang, J. Yang, Y. Chen, Y. Zhao, T.
354 Zhou, Comprehensive understanding the emission characteristics and kinetics of
355 VOCs from automotive waste paint sludge in a environmental test chamber, Journal
356 of Hazardous Materials. 429 (2022) 128387.
357 <https://doi.org/10.1016/j.jhazmat.2022.128387>.
- 358 [2] N.F. Idris, N. Le-Minh, J.E. Hayes, R.M. Stuetz, Performance of wet scrubbers to
359 remove VOCs from rubber emissions, Journal of Environmental Management. 305
360 (2022) 114426. <https://doi.org/10.1016/j.jenvman.2021.114426>.
- 361 [3] Y. Wang, H. Wang, Y. Tan, J. Liu, K. Wang, W. Ji, L. Sun, X. Yu, J. Zhao, B. Xu, J.
362 Xiong, Measurement of the key parameters of VOC emissions from wooden
363 furniture, and the impact of temperature, Atmospheric Environment. 259 (2021)
364 118510. <https://doi.org/10.1016/j.atmosenv.2021.118510>.

- 365 [4] S. Han, Q. Zhao, R. Zhang, Y. Liu, C. Li, Y. Zhang, Y. Li, S. Yin, Q. Yan, Emission
366 characteristic and environmental impact of process-based VOCs from prebaked
367 anode manufacturing industry in Zhengzhou, China, *Atmospheric Pollution*
368 *Research*. 11 (2020) 67–77. <https://doi.org/10.1016/j.apr.2019.09.016>.
- 369 [5] W. Qian, Y. Guo, X. Wang, X. Qiu, X. Ji, L. Wang, Y. Li, Quantification and
370 assessment of chemical footprint of VOCs in polyester fabric production, *Journal of*
371 *Cleaner Production*. 339 (2022) 130628.
372 <https://doi.org/10.1016/j.jclepro.2022.130628>.
- 373 [6] N. Li, Q. Jiang, F. Wang, J. Xie, Y. Li, J. Li, S. Wu, Emission behavior,
374 environmental impact and priority-controlled pollutants assessment of volatile
375 organic compounds (VOCs) during asphalt pavement construction based on
376 laboratory experiment, *Journal of Hazardous Materials*. 398 (2020).
377 <https://doi.org/10.1016/j.jhazmat.2020.122904>.
- 378 [7] B. McDonald, J. de Gouw, J. Gilman, S. Jathar, Volatile chemical products emerging
379 as largest petrochemical source of urban organic emissions, *Science*. 764 (2018)
380 760–764.
- 381 [8] H. Lee, K. Kim, Y. Choi, D. Kim, Emissions of Volatile Organic Compounds (
382 VOCs) from an Open-Circuit Dry Cleaning Machine Using a Petroleum-Based
383 Organic Solvent : Implications for Impacts on Air Quality, (2021) 4–14.
- 384 [9] F.I. Khan, A.K. Ghoshal, Removal of Volatile Organic Compounds from polluted
385 air, *Journal of Loss Prevention in the Process Industries*. 13 (2000) 527–545.

- 386 [https://doi.org/10.1016/S0950-4230\(00\)00007-3](https://doi.org/10.1016/S0950-4230(00)00007-3).
- 387 [10] X. Li, Y. Niu, H. Su, Y. Qi, Simple Thermocatalytic Oxidation Degradation of
388 VOCs, *Catalysis Letters*. (2021). <https://doi.org/10.1007/s10562-021-03770-x>.
- 389 [11] Y. Chu, J. Wang, G. Tian, R. He, Reduction in VOC emissions by intermittent
390 aeration in bioreactor landfills with gas-water joint regulation ☆, *Environmental*
391 *Pollution*. 290 (2021) 118059. <https://doi.org/10.1016/j.envpol.2021.118059>.
- 392 [12] K. Zhu, X. Wang, X. Ma, Z. Sun, X. Hu, Comparative degradation of atrazine by
393 anodic oxidation at graphite and platinum electrodes and insights into
394 electrochemical behavior of graphite anode, *Electrocatalysis*. 10 (2019) 35–44.
395 <https://doi.org/10.1007/s12678-018-0493-z>.
- 396 [13] P. Zhu, Y. Shen, L. Dai, Q. Yu, Z.-M. Zhang, C. An, Accelerating anode reaction
397 with electro-oxidation of alcohols over Ru nanoparticles to reduce the potential for
398 water splitting, *ACS Applied Materials and Interfaces*. 14 (2022) 1452–1459.
399 <https://doi.org/10.1021/acsami.1c20511>.
- 400 [14] A. Kurt, Cefuroxime Oxidation With New Generation Anodes: Evaluation Of
401 Parameter Effects, Kinetic And Total Intermediate Products, *Environmental*
402 *Research and Technology*. 4 (2021) 317–328. <https://doi.org/10.35208/ert.867139>.
- 403 [15] J. Carillo-Abad, J. Mora-Gómez, M. García-Gabaldón, E. Ortega, S. Mestre, V.
404 Pérez-Herranz, Effect of the CuO addition on a Sb-doped SnO₂ ceramic electrode
405 applied to the removal of norfloxacin in chloride media by electro-oxidation,
406 *Chemosphere*. 249 (2020) 1–9. <https://doi.org/10.1016/j.chemosphere.2020.126178>.

- 407 [16] C. Barrera-Díaz, P. Cañizares, F.J. Fernández, R. Natividad, M.A. Rodrigo,
408 Electrochemical Advanced Oxidation Processes: An Overview of the Current
409 Applications to Actual Industrial Effluents, *Journal of the Mexican Chemical*
410 *Society*. 58 (2014) 256–275.
- 411 [17] C.A. Martínez-Huitile, M.A. Rodrigo, I. Sirés, O. Scialdone, Single and Coupled
412 Electrochemical Processes and Reactors for the Abatement of Organic Water
413 Pollutants: A Critical Review, *Chemical Reviews*. 115 (2015) 13362–13407.
414 <https://doi.org/10.1021/acs.chemrev.5b00361>.
- 415 [18] M.A. Rodrigo, M.A. Oturan, N. Oturan, Electrochemically assisted remediation of
416 pesticides in soils and water: A review, *Chemical Reviews*. 114 (2014) 8720–8745.
417 <https://doi.org/10.1021/cr500077e>.
- 418 [19] I. Sirés, E. Brillas, M.A. Oturan, M.A. Rodrigo, M. Panizza, Electrochemical
419 advanced oxidation processes: Today and tomorrow. A review, *Environmental*
420 *Science and Pollution Research*. 21 (2014) 8336–8367.
421 <https://doi.org/10.1007/s11356-014-2783-1>.
- 422 [20] W.C. Cho, K.M. Poo, H.O. Mohamed, T.N. Kim, Y.S. Kim, M.H. Hwang, D.W.
423 Jung, K.J. Chae, Non-selective rapid electro-oxidation of persistent, refractory VOCs
424 in industrial wastewater using a highly catalytic and dimensionally stable Ir–Pd/Ti
425 composite electrode, *Chemosphere*. 206 (2018) 483–490.
426 <https://doi.org/10.1016/j.chemosphere.2018.05.060>.
- 427 [21] M. Tokumura, R. Nakajima, H.T. Znad, Y. Kawase, Chemical absorption process for

- 428 degradation of VOC gas using heterogeneous gas-liquid photocatalytic oxidation:
429 Toluene degradation by photo-Fenton reaction, *Chemosphere*. 73 (2008) 768–775.
430 <https://doi.org/10.1016/j.chemosphere.2008.06.021>.
- 431 [22] F. Escalona-Duran, C. Saez, J. Lobato, C. Martínez-Huitile, M. Rodrigo,
432 Electroscrubbers for removing volatile organic compounds and odorous substances
433 from polluted gaseous streams, *Current Opinion in Electrochemistry*. 28 (2021).
434 <https://doi.org/10.1016/j.coelec.2021.100718>.
- 435 [23] P. Cañizares, C. Sáez, J. Lobato, R. Paz, M.A. Rodrigo, Effect of the Operating
436 Conditions on the Oxidation Mechanisms in Conductive-Diamond Electrolyses,
437 *Journal of The Electrochemical Society*. 154 (2007) E37.
438 <https://doi.org/10.1149/1.2424409>.
- 439 [24] M. Govindan, R. Adam Gopal, I.S. Moon, Efficient removal of gaseous
440 trichloroethylene by continuous feed electro-scrubbing using a new homogeneous
441 heterobimetallic electro-catalyst, *Chemical Engineering Journal*. 308 (2017) 1145–
442 1153. <https://doi.org/10.1016/j.cej.2016.09.137>.
- 443 [25] F. Escalona-Durán, M. Muñoz-Morales, C.M. Fernández-Marchante, J. Lobato, C.A.
444 Martínez-Huitile, M.A. Rodrigo, Modelling electro-scrubbers for removal of VOCs,
445 *Separation and Purification Technology*. 277 (2021).
446 <https://doi.org/10.1016/j.seppur.2021.119419>.
- 447 [26] M. Castañeda-Juárez, M. Muñoz-Morales, F.L. Souza, C. Sáez, P. Cañizares, P.T.
448 Almazán-Sánchez, I. Linares-Hernández, M.A. Rodrigo, Electro-absorbers: A

- 449 comparison on their performance with jet-absorbers and absorption columns,
450 Catalysts. 10 (2020) 1–14. <https://doi.org/10.3390/catal10060653>.
- 451 [27] S.J. Chung, I.S. Moon, An improved method of removal for high concentrations of
452 NO by electro-scrubbing process, Process Safety and Environmental Protection. 91
453 (2013) 153–158. <https://doi.org/10.1016/j.psep.2012.05.010>.
- 454 [28] M. Govindan, R. Adam Gopal, I.S. Moon, Efficient removal of gaseous
455 trichloroethylene by continuous feed electro-scrubbing using a new homogeneous
456 heterobimetallic electro-catalyst, Chemical Engineering Journal. 308 (2017) 1145–
457 1153. <https://doi.org/10.1016/j.cej.2016.09.137>.
- 458 [29] F. Escalona-Duran, M. Muñoz-Morales, K. de Freitas Araújo, C. Saez, P. Cañizares,
459 C. Martinez-Huitle, M. Rodrigo, Comparison of the performance of packed column
460 and jet electro-scrubbers for the removal of toluene, Journal of Environmental
461 Chemical Engineering. 9 (2021). <https://doi.org/10.1016/j.jece.2021.106114>.
- 462 [30] K. Chandrasekara Pillai, S.J. Chung, T. Raju, I.S. Moon, Experimental aspects of
463 combined NO_x and SO₂ removal from flue-gas mixture in an integrated wet
464 scrubber-electrochemical cell system, Chemosphere. 76 (2009) 657–664.
465 <https://doi.org/10.1016/j.chemosphere.2009.04.013>.
- 466 [31] S. Ardizzone, C.L. Bianchi, G. Cappelletti, A. Naldoni, C. Pirola, Photocatalytic
467 degradation of toluene in the gas phase: Relationship between surface species and
468 catalyst features, Environmental Science and Technology. 42 (2008) 6671–6676.
469 <https://doi.org/10.1021/es8009327>.

- 470 [32] J. Al-Sabahi, T. Bora, M. Al-Abri, J. Dutta, Efficient visible light photocatalysis of
471 benzene, toluene, ethylbenzene and xylene (BTEX) in aqueous solutions using
472 supported zinc oxide nanorods, PLoS ONE. 12 (2017) 1–16.
473 <https://doi.org/10.1371/journal.pone.0189276>.
- 474 [33] A.N. Arias, R. de Mello, J. Lobato, A.J. Motheo, M.A. Rodrigo, Electrolytic removal
475 of volatile organic compounds: Keys to understand the process, Journal of
476 Electroanalytical Chemistry. 912 (2022) 116259.
477 <https://doi.org/10.1016/j.jelechem.2022.116259>.
- 478 [34] T. Goto, M. Ogawa, Efficient photocatalytic oxidation of benzene to phenol by metal
479 complex-clay/TiO₂ hybrid photocatalyst, RSC Advances. 6 (2016) 23794–23797.
480 <https://doi.org/10.1039/c5ra25430b>.
- 481 [35] I. Banga, A. Paul, A.U. Sardesai, S. Muthukumar, S. Prasad, M.A.T.H: Methanol
482 vapor analytics through handheld sensing platform, Electrochimica Acta. 368 (2021)
483 137624. <https://doi.org/10.1016/j.electacta.2020.137624>.
- 484 [36] H. Cerfontain, Solubility of aromatic hydrocarbons in aqueous sulfuric acid, Recueil
485 Des Travaux Chimiques Des Pays-Bas. 84 (1965) 491–502.
486 <https://doi.org/10.1002/recl.19650840414>.
- 487 [37] M. Govindan, S.J. Chung, H.H. Moon, J.W. Jang, I.S. Moon, Development of a
488 biphasic electroreactor with a wet scrubbing system for the removal of gaseous
489 Benzene, ACS Combinatorial Science. 15 (2013) 439–446.
490 <https://doi.org/10.1021/co400046g>.

- 491 [38] K.-W. Kim, M. Kuppuswamy, R.F. Savinell, Electrochemical oxidation of benzene
492 at a glassy carbon electrode, *Journal of Applied Electrochemistry*. 30 (2000) 543–
493 549.
- 494 [39] R.T.S. Oliveira, G.R. Salazar-Banda, M.C. Santos, M.L. Calegari, D.W. Miwa,
495 S.A.S. Machado, L.A. Avaca, Electrochemical oxidation of benzene on boron-doped
496 diamond electrodes, *Chemosphere*. 66 (2007) 2152–2158.
497 <https://doi.org/10.1016/j.chemosphere.2006.09.024>.
- 498 [40] I. Cesarino, V. Cesarino, F.C. Moraes, T.C.R. Ferreira, M.R.V. Lanza, L.H.
499 Mascaro, S.A.S. Machado, Electrochemical degradation of benzene in natural water
500 using silver nanoparticle-decorated carbon nanotubes, *Materials Chemistry and*
501 *Physics*. 141 (2013) 304–309. <https://doi.org/10.1016/j.matchemphys.2013.05.015>.
- 502 [41] J. Schlauer, Absorption, 1. Fundamentals, in: *Ullmann's Encyclopedia of Industrial*
503 *Chemistry*, Wiley-VCH Verlag GmbH & Co. KGaA, 2008.
504 https://doi.org/10.1002/14356007.b03_08.pub2.
- 505 [42] A.G. Ramu, G. Muthuraman, P. Silambarasan, M. Il Shik, Sustainable generation of
506 homogeneous Fe(VI) oxidant for the room temperature removal of gaseous N₂O by
507 electro-scrubbing process, *Chemosphere*. 272 (2021) 129497.
508 <https://doi.org/10.1016/j.chemosphere.2020.129497>.
- 509 [43] M. Govindan, C. Sang-Joon, M. Il-Shik, Mineralization of Gaseous Acetaldehyde by
510 Electrochemically Generated Co(III) in H₂SO₄ with Wet Scrubber Combinatorial
511 System, *ACS Combinatorial Science*. 14 (2012) 359–365.

512 [44] M. Govindan, I.S. Moon, A single catalyst of aqueous CoIII for deodorization of
513 mixture odor gases: A development and reaction pathway study at electro-scrubbing
514 process, *Journal of Hazardous Materials*. 260 (2013) 1064–1072.
515 <https://doi.org/10.1016/j.jhazmat.2013.06.055>.

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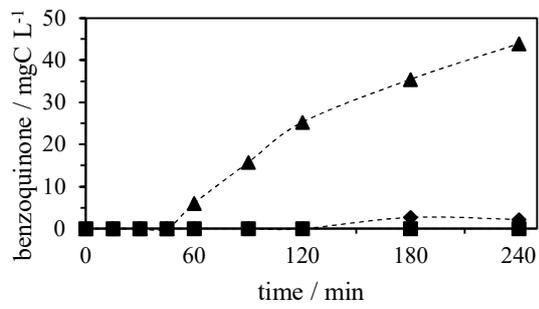
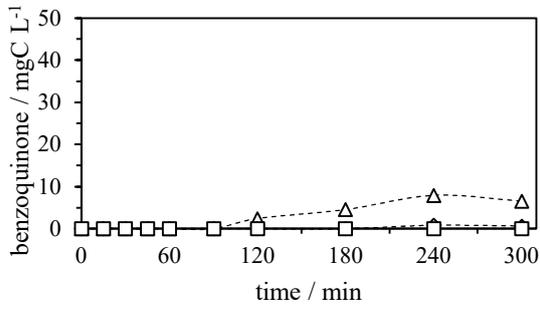
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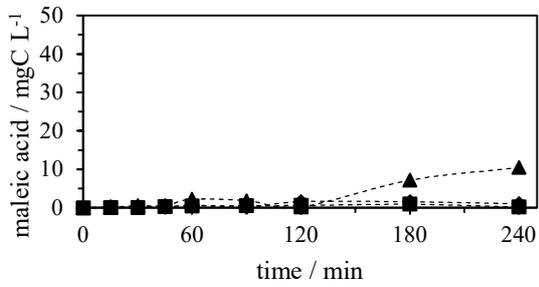
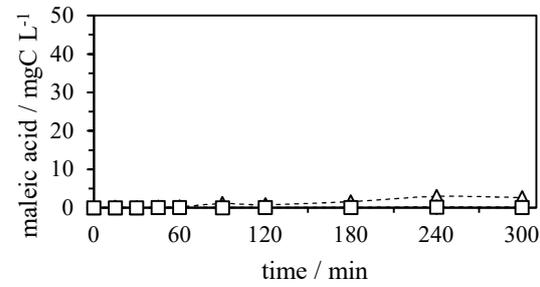
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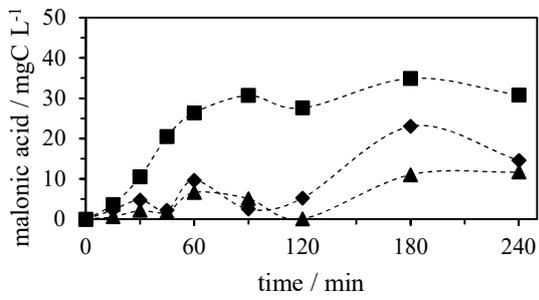
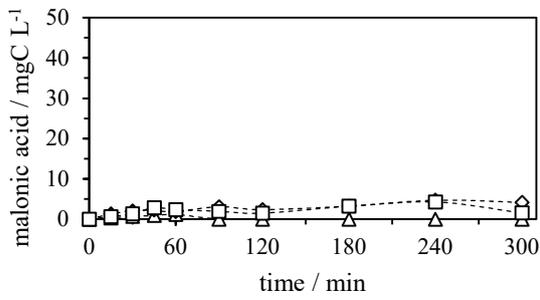
Appendix I: Supplementary material



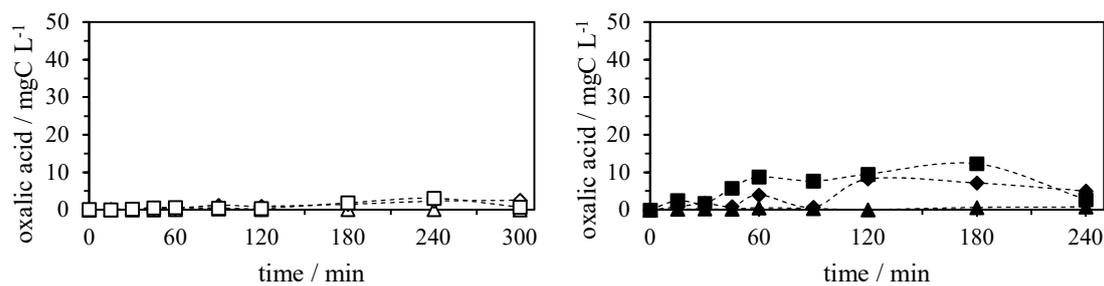
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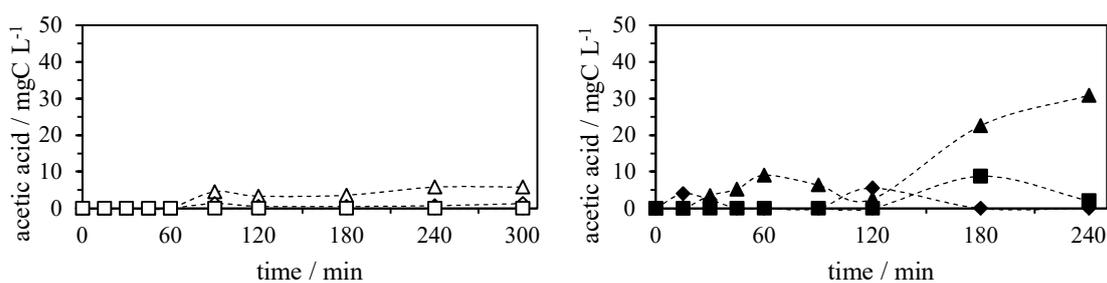
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536 Figure S1. Time evolution of each intermediates compounds identified in the liquid phase at
 537 different gaseous flow of 1.5 (white) and 6 (black) L h⁻¹ in electro-absorption processes at 10
 538 (triangle), 50 (rhombus) and 100 (square) mA cm⁻².

539

540 Table S1. Mass of benzene (mgC) retained on the active carbon after the electro-absorption
 541 process and the percentage of each value in relation to the total mass entering the system.

Inlet gas flow (L h ⁻¹)	Current density (mA cm ⁻²)			
	0	10	50	100
1.5	708.52 (57%)	355.88 (29%)	246.23 (19%)	137.37 (11%)
6.0	3208.86 (68%)	2858.70 (63%)	2463.70 (52%)	1940.83 (44%)

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